



EPA United States
Environmental
Protection Agency

Draft Plan for Review of the Secondary National Ambient Air Quality Standards for Nitrogen Dioxide and Sulfur Dioxide

Notice

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Draft Plan for Review of the Secondary National Ambient Air Quality Standards for Nitrogen Dioxide and Sulfur Dioxide

U. S. Environmental Protection Agency
National Center for Environment Assessment
Office of Research and Development
and
Office of Air Quality Planning and Standards
Office of Air and Radiation
Research Triangle Park, North Carolina 27711

DISCLAIMER

This draft plan for the review of the secondary national ambient air quality standards for nitrogen dioxide and sulfur dioxide is an informational document prepared for external review purposes and does not constitute U.S. Environmental Protection Agency policy. This plan also serves as a management tool for the U.S. Environmental Protection Agency's National Center for Environmental Assessment and the Office of Air Quality Planning and Standards in Research Triangle Park, North Carolina. This information may be modified to reflect information developed during this review and to address advice and comments received from the Clean Air Scientific Advisory Committee and the public throughout this review. Mention of trade names or commercial products does not constitute endorsement or recommendation for use.

Key Terms

Acidification: The process of increasing the acidity of a system (lake, stream, forest soil). Atmospheric deposition of acidic or acidifying compounds can acidify lakes, streams and forest soils.

Air Quality Indicator: The substance or set of substances (e.g., PM_{2.5}, NO₂, SO₂) occurring in the ambient air for which a level is set for a NAAQS standard and monitoring occurs.

ANC (Acid Neutralizing Capacity): A key indicator of the ability of water to neutralize the acid or acidifying inputs it receives. This ability depends largely on associated biogeophysical characteristics.

Biologically Relevant Indicator: A stressor to which organisms respond.

Dry Deposition: The removal of gases and particles from the atmosphere to surfaces in the absence of precipitation (rain, snow) or occult deposition.

Ecosystem: A dynamic complex of interacting plants, animals, and microorganisms and the non-living environment.

Ecosystem Benefit: The value, expressed either qualitatively, quantitatively and/or in economic terms where possible, associated with changes in ecosystem services that result, either directly or indirectly in improved human health and/or welfare. Some examples of ecosystem benefits that derive from improved air quality include improvements in habitat for sport fish species, drinking water quality, visual quality of scenic views, and quality of recreational areas.

Ecosystem Services: Conditions and processes through which natural ecosystems, and the species that are part of them, currently help sustain and fulfill human life or have the potential to do so in the future. Examples include, clean air, clean water, food and fiber production, flood protection, water purification, pollination, and pest control. Improvements in these services may be valued as an ecosystem benefit.

Eutrophication: The process by which a body of water becomes enriched in dissolved nutrients that stimulate the growth of aquatic plant life usually resulting in the depletion of dissolved oxygen

Occult Deposition: The removal of gases and particles from the atmosphere to surfaces by fog or mist.

Welfare Effects: Effects on soils, water, crops, vegetation, man-made materials, animals, wildlife, weather, visibility and climate, damage to and deterioration of property,

and hazards to transportation, as well as effects on economic values and on personal comfort and well-being, whether caused by transformation, conversion, or combination with other air pollutants (CAA 302(h))

Wet Deposition: The removal of gases and particles from the atmosphere to surfaces by rain or other precipitation

Valuation: The economic or non-economic process of determining either the value of maintaining a given ecosystem type, state or condition, or the value of a change in, an ecosystem, its components, or the services it provides.

TABLE OF CONTENTS

1.	INTRODUCTION.....	1-1
1.1	LEGISLATIVE REQUIREMENTS	1-2
1.2	REGULATORY HISTORY OF THE SECONDARY NAAQS FOR NO ₂ AND SO ₂	1-3
1.3	OVERVIEW OF THE NAAQS REVIEW PROCESS.....	1-7
1.4	SCIENCE PRIMER: NO _x /SO _x EFFECTS ON ECOSYSTEMS.....	1-9
2.	REVIEW SCHEDULE.....	2-1
3.	KEY POLICY-RELEVANT ISSUES	3-1
3.1	ISSUES TO BE CONSIDERED IN THE CURRENT REVIEW.....	3-1
4.	SCIENCE ASSESSMENT.....	4-1
4.1	SCOPE AND ORGANIZATION	4-1
4.2	ASSESSMENT APPROACH.....	4-2
5.	RISK/EXPOSURE ASSESSMENT.....	5-1
5.1	ASSESSMENT APPROACH.....	5-2
5.2	TOOLS FOR RISK ASSESSMENT	5-4
5.3	KNOWLEDGE GAPS.....	5-4
5.3	PUBLIC AND SCIENTIFIC REVIEW	5-6
6.	POLICY ASSESSMENT.....	6-1
7.	REFERENCES.....	7-1
8.	APPENDIX A	A-1
9.	APPENDIX B	B-1

1. INTRODUCTION

The U.S. Environmental Protection Agency (EPA) is conducting a review of the existing primary (health-based) and secondary (welfare-based) National Ambient Air Quality Standards (NAAQS) for nitrogen dioxide (NO₂) and sulfur dioxide (SO₂). The reviews of the primary NAAQS for NO₂ and for SO₂ are addressed in separate plans released during the winter of 2006/2007. The purpose of this document is to communicate the draft integrated plan for the joint review of the secondary NAAQS for these pollutants. In this document, the terms NO₂ and NO_x and SO₂ and SO_x are not interchangeable. The terms NO_x (oxides of nitrogen) and SO_x (sulfur oxides) refer to the listed Criteria Air Pollutants for which EPA has regulatory authority under Sections 108 and 109 of the Clean Air Act (CAA), and for which criteria must be developed and reviewed every 5 years. The terms NO₂ and SO₂ refer to the specific air quality indicators (pollutant species) specified by the current standards whose concentrations are monitored to determine whether the NAAQS is being met in a given location. Section 1.1 below includes additional explanation on the use of these terms within the context of the various phases of the new NAAQS review process. This review will evaluate new information published in the peer-reviewed literature since the completion of the last NO₂ (1995) and SO₂ (1996) reviews, including assessments of the adequacy of the current secondary NAAQS, and consider the possible need for a new single indicator or suite of indicators, as well as changed or retained level(s) and/or averaging times for the standards, which may include nitrogen and sulfur compounds other than NO₂ and SO₂.

This review plan is organized into six chapters. Chapter 1 presents background information on the recently-revised NAAQS review process; on the nature of the NO_x/SO_x problem; on the legislative requirements for the review of the NAAQS; and on the regulatory history of past reviews of the NAAQS for NO₂ and SO₂. Chapter 2 presents the proposed review schedule. Chapter 3 presents a set of key policy-relevant questions that will serve to focus the NAAQS review process on the critical scientific and policy issues. Chapters 4 through 6 discuss the science, risk/exposure, and policy assessment portions of the review. As the assessments proceed, the plan described here may be modified to reflect information received during the review process and to address advice and comments received from the Clean Air Scientific Advisory Committee and from the public throughout this review.

1.1 LEGISLATIVE REQUIREMENTS

Two sections of the Clean Air Act (CAA) govern the establishment and revision of the NAAQS. Section 108 (42 U.S.C. 7408) directs the Administrator to identify and list “air pollutants” that “in his judgment, may reasonably be anticipated to endanger public health and welfare” and whose “presence . . . in the ambient air results from numerous or diverse mobile or stationary sources” and to issue air quality criteria for those that are listed. Air quality criteria are intended to “accurately reflect the latest scientific knowledge useful in indicating the kind and extent of identifiable effects on public health or welfare which may be expected from the presence of [a] pollutant in ambient air”

Section 109 (42 U.S.C. 7409) directs the Administrator to propose and promulgate “primary” and “secondary” NAAQS for pollutants listed under section 108. A secondary standard, as defined in Section 109(b)(2), must “specify a level of air quality the attainment and maintenance of which, in the judgment of the Administrator, based on such criteria, is required to protect the public welfare from any known or anticipated adverse effects associated with the presence of [the] pollutant in the ambient air.”¹

In setting standards that are “requisite” to protect public health and welfare, as provided in section 109(b), EPA’s task is to establish standards that are neither more nor less stringent than necessary for these purposes. In so doing, EPA may not consider the costs of implementing the standards. (See generally *Whitman v. American Trucking Associations*, 531 U.S. 457, 465-472, 475-76 (2001).)

Section 109(d)(1) requires that “not later than December 31, 1980, and at 5-year intervals thereafter, the Administrator shall complete a thorough review of the criteria published under section 108 and the national ambient air quality standards . . . and shall make such revisions in such criteria and standards and promulgate such new standards as may be appropriate”

Section 109(d)(2) requires that an independent scientific review committee “shall complete a review of the criteria . . . and the national primary and secondary ambient air quality standards . .

¹ Welfare effects as defined in section 302(h) [42 U.S.C. 7602(h)] include, but are not limited to, “effects on soils, water, crops, vegetation, man-made materials, animals, wildlife, weather, visibility and climate, damage to and deterioration of property, and hazards to transportation, as well as effects on economic values and on personal comfort and well-being.”

1 . and shall recommend to the Administrator any new . . . standards and revisions of existing
2 criteria and standards as may be appropriate . . .” Since the early 1980's, this independent
3 review function has been performed by the Clean Air Scientific Advisory Committee (CASAC)
4 of EPA’s Science Advisory Board.
5

6 **1.2 REGULATORY HISTORY OF THE SECONDARY NAAQS FOR NO₂** 7 **AND SO₂**

8 **1.2.1 NO₂ NAAQS.**

9 In 1971, the first air quality criteria document for NO_x was issued by the National Air
10 Pollution Control Association (NAPCA), one of EPA’s predecessor agencies. After reviewing
11 the relevant science on the public health and welfare effects associated with oxides of nitrogen,
12 EPA promulgated identical primary and secondary NAAQS for NO₂ on April 30, 1971. Under
13 section 109 of the Act, these standards were set at 0.053 parts per million (ppm) as an annual
14 average (36 FR 8186). In 1982, EPA published *Air Quality Criteria for Oxides of Nitrogen*
15 (EPA, 1982), which updated the scientific criteria upon which the initial standards were based.
16 On February 23, 1984, EPA proposed to retain these standards (49 FR 6866). After taking into
17 account public comments, EPA published the final decision to retain these standards on June 19,
18 1985 (50 FR 25532).

19 In November 1991, EPA released an updated draft Air Quality Criteria Document
20 (AQCD) for CASAC and public review and comment (56 FR 59285). The draft AQCD
21 provided a comprehensive assessment of the available scientific and technical information on
22 health and welfare effects associated with NO₂ and other NO_x compounds. The CASAC
23 reviewed the document at a meeting held on July 1, 1993 and concluded in a closure letter to the
24 Administrator that the document “provides a scientifically balanced and defensible summary of
25 current knowledge of the effects of this pollutant and provides an adequate basis for EPA to
26 make a decision as to the appropriate NAAQS for NO₂” (Wolff, 1993).

27 The EPA also prepared a draft Staff Paper that summarized and integrated the key studies
28 and scientific evidence contained in the revised AQCD and identified the critical elements to be
29 considered in the review of the NO₂ NAAQS. The Staff Paper received external review at a
30 December 12, 1994 CASAC meeting. The CASAC comments and recommendations were
31 reviewed by EPA staff and incorporated into the final draft of the Staff Paper, as appropriate.
32 The CASAC reviewed the final draft of the Staff Paper in June, 1995 and responded by written

1 closure letter (Wolff, 1995). In September, 1995, EPA finalized the Staff Paper, “Review of the
2 National Ambient Air Quality Standards for Nitrogen Dioxide: Assessment of Scientific and
3 Technical Information,” (EPA, 1995).

4 On October 2, 1995, the Administrator announced her proposed decision not to revise
5 either the primary or secondary NAAQS for NO₂ based on the information available to her in
6 this review (60 FR 52874; October 11, 1995). After careful evaluation of the comments received
7 on the October 1995 proposal, the Administrator made a final determination that revisions to
8 neither the primary nor the secondary NAAQS for NO₂ were appropriate at that time (61 FR
9 52852, October 8, 1996). The level for both the existing primary and secondary NAAQS for
10 NO₂ remains 0.053 ppm (equivalent to 100 micrograms per cubic meter of air [ug/m³]) in annual
11 arithmetic average, calculated as the arithmetic mean of the 1-hour NO₂ concentrations.
12

13 **1.2.2 SO₂ NAAQS.**

14 Based on the 1970 sulfur oxides criteria document (DHEW, 1970), EPA promulgated
15 primary and secondary NAAQS for SO₂, under section 109 of the Act on April 30, 1971 (36 FR
16 8186). The secondary standards included a standard at 0.02 ppm in an annual arithmetic mean
17 and a 3-hr average of 0.5 ppm, not to be exceeded more than once per year. These secondary
18 standards were established solely on the basis of vegetation effects evidence. In 1973, revisions
19 made to Chapter 5 “Effects of Sulfur Oxide in the Atmosphere on Vegetation” of Air Quality
20 Criteria for Sulfur Oxides (EPA 1973), indicated that it could not properly be concluded that the
21 vegetation injury reported resulted from the average SO₂ exposure over the growing season,
22 rather than the short-term peak concentrations. EPA, therefore, proposed (38 FR 11355) and
23 then finalized a revocation of the annual mean secondary standard (38 FR 25678). At that time,
24 EPA was aware that sulfur oxides have other public welfare effects, including effects on
25 materials, visibility, soils and water. However, the available data were considered insufficient to
26 establish a quantitative relationship between specific sulfur dioxide concentrations and effects
27 needed for standard setting (38 FR 25679).

28 In 1979, EPA announced that it was revising the AQCD for sulfur oxides concurrently
29 with that for particulate matter and would produce a combined particulate matter (PM) and sulfur
30 oxides criteria document. Following its review of a draft revised criteria document in August,
31 1980, the CASAC concluded that acidic deposition was a topic of extreme scientific complexity
32 because of the difficulty in establishing firm quantitative relationships among: (1) emissions of
September 2007

1 relevant pollutants (e.g., SO₂ and oxides of nitrogen), (2) formation of acidic wet and dry
2 deposition products, and (3) effects on terrestrial and aquatic ecosystems. CASAC also noted
3 that acidic deposition involves, at a minimum, several different criteria pollutants - oxides of
4 sulfur, oxides of nitrogen, and the fine particulate fraction of suspended particles. The
5 Committee felt that any document on this subject should address both wet and dry deposition,
6 since dry deposition was believed to account for at least one half of the total acid deposition
7 problem.

8 For these reasons, the CASAC recommended that a separate, comprehensive document on
9 acidic deposition be prepared prior to any consideration of using the NAAQS as a regulatory
10 mechanism for the control of acidic deposition. CASAC also suggested that a discussion of
11 acidic deposition be included in the AQCDs for both nitrogen oxides and PM and SO_x. In
12 response to these recommendations, EPA subsequently prepared the following documents: The
13 Acidic Deposition Phenomenon and Its Effects: Critical Assessment Review Papers, Volumes I
14 and II (EPA, 1984), and The Acidic Deposition Phenomenon and Its Effects: Critical Assessment
15 Document (EPA, 1985) (53 FR 14935 -14936). These documents, though they were not
16 considered criteria documents and did not undergo CASAC review, represented the most
17 comprehensive summary of relevant scientific information completed by the EPA at that point.

18 Following CASAC closure on the criteria document for SO₂ in December 1981, EPA
19 OAQPS published a Staff Paper in November, 1982. The issue of acidic deposition was not,
20 however, assessed directly in this Staff Paper because EPA followed the guidance given by
21 CASAC.

22 On April 26, 1988 (53 FR 14926), EPA proposed not to revise the existing primary and
23 secondary standards. This proposal regarding the secondary SO₂ NAAQS was due to the
24 Administrators conclusions that (1) based upon the then-current scientific understanding of the
25 acidic deposition problem, it would be premature and unwise to prescribe any regulatory control
26 program at that time, and (2) when the fundamental scientific uncertainties had been reduced
27 through ongoing research efforts, EPA would draft and support an appropriate set of control
28 measures.

29 In spite of the complexities and significant remaining uncertainties associated with the
30 acidic deposition problem, it soon became clear that a program to address acidic deposition was
31 needed. On November 15, 1990, Amendments to the CAA were passed by Congress and signed
32 into law by the President. In Title IV of these Amendments, Congress included a statement of

1 findings that had led them to take this action, including that: (1) the presence of acidic
2 compounds and their precursors in the atmosphere and in deposition from the atmosphere
3 represents a threat to natural resources, ecosystems, materials, visibility, and public health; (2)
4 the problem of acid deposition is of national and international significance; and (3) current and
5 future generations of Americans will be adversely affected by delaying measures to remedy the
6 problem...”. The goal of Title IV was to reduce emissions of SO₂ by 10 million tons and NO_x
7 emissions by 2 million tons from 1980 emission levels in order to achieve reductions over broad
8 geographic regions/areas.

9 Congress, however, clearly envisioned that further action might be necessary in the long
10 term and reserved judgment on the form it could take, as evidenced by the inclusion of section
11 404 of the 1990 Amendments (Clean Air Act Amendments of 1990, Pub. L. 101-549, § 404).
12 This section required EPA to conduct a study on the feasibility and effectiveness of an acid
13 deposition standard or standards to protect “sensitive and critically sensitive aquatic and
14 terrestrial resources” and at the conclusion of the study, submit a report to Congress. Five years
15 later EPA submitted to Congress its report titled Acid Deposition Standard Feasibility Study:
16 Report to Congress (EPA, 1995) in fulfillment of this requirement. The Acid Deposition
17 Standard Feasibility Study Report to Congress concluded that establishing acid deposition
18 standards for sulfur and nitrogen deposition may at some point in the future be technically
19 feasible although appropriate deposition loads for these acidifying chemicals could not defined
20 with reasonable certainty at that time.

21 The 1990 Amendments also added new language to sections of the CAA pertaining to the
22 scope or application of the secondary NAAQS designed to protect the public welfare. Section
23 108 (g) specified that “the Administrator may assess the risks to ecosystems from exposure to
24 Criteria Air Pollutants (as identified by the Administrator in the Administrator’s sole
25 discretion)”. The definition of public welfare in section 302 (h) was expanded to state that the
26 welfare effects identified should be protected from adverse effects associated with criteria air
27 pollutants “...whether caused by transformation, conversion, or combination with other air
28 pollutants.”

29 In 1999, seven Northeastern States cited this amended language in section 302 (h) in a
30 petition to EPA to use its authority under the NAAQS program to promulgate secondary
31 NAAQS for the criteria pollutants associated with the formation of acid rain. The petition stated
32 that this language “clearly references the transformation of pollutants resulting in the inevitable

1 formation of sulfate and nitrate aerosols and/or their ultimate environmental impacts as wet and
2 dry deposition, clearly signaling Congressional intent that the welfare damage occasioned by
3 sulfur and nitrogen oxides be addressed through the secondary standard provisions of Section
4 109 of the Act. The petition further stated that “recent federal studies, including the NAPAP
5 Biennial Report to Congress: An Integrated Assessment, document the continued-and increasing-
6 damage being inflicted by acid deposition to the lakes and forests of New York, New England
7 and other parts of our nation, demonstrating that the Title IV program had proven insufficient.”
8 The petition also listed other adverse welfare effects associated with the transformation of these
9 criteria pollutants, including visibility impairment, eutrophication of coastal estuaries, global
10 warming, tropospheric ozone and stratospheric ozone depletion.

11 In a related matter, the U.S. Department of Interior (DOI) requested in 2000 that the EPA
12 initiate a rulemaking proceeding to enhance the air quality in national parks and wilderness areas
13 in order to protect resources and values that are being adversely affected by air pollution.
14 Included among the effects of concern identified in the request were acidification of streams,
15 surface waters and/or soils, eutrophication of coastal waters, visibility impairment, and foliar
16 injury from ozone.

17 In a Federal Register notice in 2001, EPA announced receipt of these items and requested
18 comment on the issues raised by these requests. EPA stated that it would consider any relevant
19 comments and information submitted, along with the information provided by the petitioners and
20 DOI, before making any decision concerning a response to these requests for rulemaking.

21 In this review, EPA will again revisit the appropriateness and feasibility of setting a
22 secondary NAAQS to address the welfare effects resulting from the deposition of these criteria
23 pollutants and their transformation products. This draft plan describes potential elements of that
24 review.

26 **1.3 OVERVIEW OF THE NAAQS REVIEW PROCESS**

27 U.S. EPA has recently decided to make a number of changes to the process for reviewing
28 the NAAQS (described at [www.http://epa.gov/ttn/naaqs/](http://epa.gov/ttn/naaqs/)). The revised NAAQS review process
29 contains four major components: an integrated review plan, a science assessment, a risk/exposure
30 assessment, and a policy assessment/rulemaking. In addition to these procedural modifications,
31 for this secondary NAAQS review we have decided to examine two of the six criteria pollutants,
32 oxides of nitrogen and sulfur oxides, together, rather than individually, as has been done in the

1 past. This decision derives from the fact that NO_x, SO_x, and their associated transformation
2 products are linked from an atmospheric chemistry perspective, as well as from an environmental
3 effects perspective (most notably in the case of secondary aerosol formation and acidification in
4 ecosystems). These interactions have been recognized historically by both CASAC and EPA;
5 the science related to these interactions continues to evolve and grow, emphasizing the
6 importance of considering them together.

7 The first phase of the revised NAAQS review process is the development of the integrated
8 review plan. This document represents the current draft plan and specifies the schedule of the
9 review, the process for conducting the review, and the key policy-relevant science issues that
10 will guide the review. This draft plan will be submitted for review and comment to CASAC,
11 other non-EPA scientists, and the public. The final integrated review plan will take into account
12 comments from these entities.

13 The second phase of the process is the development of the science assessment, which
14 consists of the Integrated Science Assessment (ISA) for NO_x and SO_x and supporting details in
15 its annexes. The U.S. EPA's National Center for Environmental Assessment (NCEA) along with
16 contractor support is currently developing the annexes to the ISA. These annexes will contain a
17 comprehensive description and evaluation/assessment of the full breadth of the recent scientific
18 literature pertaining to known and anticipated effects on public welfare associated with the
19 presence of the NO_x and SO_x criteria pollutant(s) in the ambient air, emphasizing the
20 information that has become available since the last review in order to reflect the current state of
21 knowledge. NCEA will then critically evaluate, integrate, and synthesize the most policy-
22 relevant science from the annexes into an ISA. The ISA is intended to provide information useful
23 in forming judgments about air quality indicator(s), form(s), averaging time(s) and level(s) for
24 the secondary NAAQS. Hence, the ISA and its associated annexes function in the new NAAQS
25 review in part as the Air Quality Criteria Document (AQCD) did in previous reviews. The
26 schedule includes production of a first and second draft ISA, both of which will undergo CASAC
27 and public review prior to completion of the final ISA. Section 4 provides a more detailed
28 description of the planned scope, organization and assessment approach for the annexes and ISA,
29 as well as EPA's rationale for conducting a joint review of these criteria pollutants.

30 In the third phase of the revised review process, the risk/exposure assessment, OAQPS
31 plans to draw upon the ISA to develop quantitative and qualitative estimates of the risks of
32 adverse welfare effects occurring as a result of current ambient levels of nitrogen oxides and

sulfur oxides, levels that meet the current standards for NO₂ and SO₂, or levels that meet possible alternative standards. Section 5 of this draft Plan contains more detail about possible approaches EPA could take in conducting the risk/exposure Assessment. Once the draft ISA is complete, EPA will release a Scope and Methods Plan/document for CASAC and public review that describes the actual scope of the analyses to be performed and the tools/methods that will be employed. Once the risk/exposure assessment is complete, a report will be developed that focuses on key results, observations, and uncertainties and may include results from specific analyses designed to inform the review. This risk/exposure assessment report will also undergo review by CASAC and the public.

The fourth component of the revised process will be a policy assessment/rulemaking. Under the revised NAAQS process, a policy assessment reflecting Agency views will be published in the Federal Register as an Advanced Notice of Proposed Rulemaking (ANPR). The ANPR will be accompanied by supporting documents, such as air quality analyses and technical support documents, as appropriate. The ANPR takes the place of the Staff Paper prepared for previous NAAQS reviews. Issuance of a Proposed and Final Rule will then complete the rulemaking process.

1.4 SCIENCE PRIMER: NO_x/SO_x EFFECTS ON ECOSYSTEMS

The following section is intended to provide a brief summary of the generally-accepted effects of NO_x and SO_x pollution on ecosystems. The information presented here is drawn from the last NO_x (1995) and SO_x (1996) NAAQS reviews as well as the generally accepted science to date.

Emissions of NO_x and SO_x compounds into the air react through a complex series of gas-phase and heterogeneous reactions to produce additional intermediate compounds and final products. These reactions with NO_x and SO_x most often occur in the same airmasses and under the meteorological influence as those acting on formation of ozone (O₃) and secondary aerosols. These nitrogen- and sulfur-containing compounds are removed from the air by deposition – wet (rain, snow), occult (fog, mist), or dry (gases and particles) -- onto surfaces. Prevailing winds can transport these compounds hundreds of miles and across state and national borders.

Deposition of chemical species derived from NO_x and SO_x emissions to the environment initiates changes in ecosystem biochemistry, structure, and function. Since both NO_x and SO_x emissions can react in the atmosphere to form strong acids, one of the possible environmental endpoints is

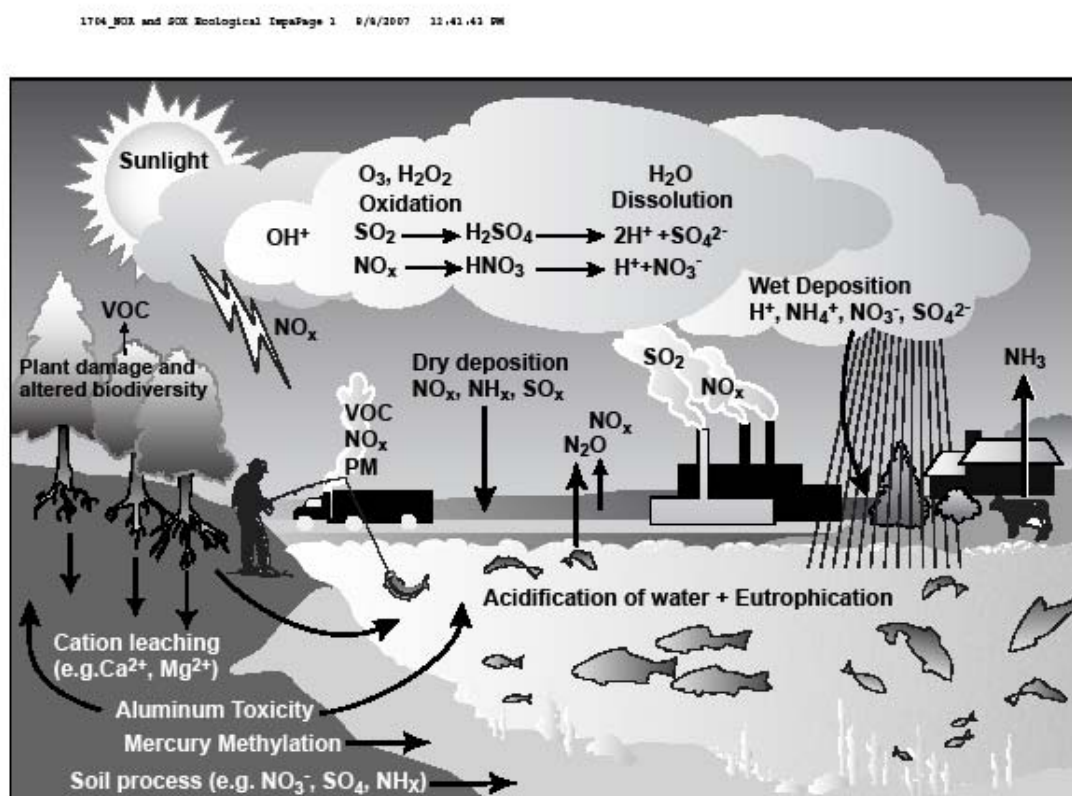
1 acidification. Acidification results in a cascade of effects that alter biogeochemical cycles and harm
2 terrestrial and aquatic ecosystems, effects including slower growth, the injury or death of forests, and
3 localized extinction of fish and other aquatic species. Acidic deposition can also cause accelerated
4 erosion of exposed materials and structures, such as buildings and statues.

5 In addition to acidification, NO_x acts with other sources of reactive nitrogen, such as ammonia-
6 based nitrogen from synthetic fertilizer and animal waste, to cause “Nitrogen pollution”, resulting in a
7 suite of terrestrial and aquatic ecological problems including biodiversity losses, community shifts,
8 eutrophication, disease, and harmful algal blooms.

9 Lastly, SO_x interacts with mercury (Hg) in ecosystems to increase the production of
10 methylmercury, a powerful toxin that bioaccumulates, often to the point of causing toxic doses to
11 top members of food chains (e.g., river otters and panthers).

12 A summary illustration of NO_x and SO_x effects on the environment is presented by
13 Figure 1.1.

Figure 1.1 Nitrogen and Sulfur cycling and interactions in the environment.



1.4.1 Acidification Effects on Freshwater and Terrestrial Ecosystems from Sulfur and Nitrogen Deposition

The process of acidification affects both freshwater aquatic and terrestrial ecosystems. Acid deposition causes acidification of sensitive surface waters. The effects of acidic deposition on aquatic systems depend largely upon the ability of the ecosystem to neutralize the additional acid. As acidity increases, aluminum leached from soils and sediments flows into lakes and streams and can be toxic to aquatic species. The lower pH levels and higher aluminum levels

1 resulting from acidification make it difficult for some fish and other aquatic species to survive,
2 grow, and reproduce.

3 Effects of acidic deposition on forest ecosystems has come to focus increasingly on the
4 biogeochemical processes that affect uptake, retention, and cycling of nutrients within these
5 ecosystems. Decreases in base cations from soils are at least partly attributable to acidic
6 deposition. Base cation depletion is a cause for concern because of the role these ions play in
7 acid neutralization and, in the case of calcium, magnesium and potassium, are essential nutrients
8 for plant growth and physiology. Changes in the relative proportions of these nutrients have
9 been correlated with declining forest health.

10 11 **1.4.2 Excess Nitrogen in terrestrial and estuarine ecosystems**

12 In addition to acidification effects, NO_x contributes to total nitrogen loading of
13 ecosystems. In terrestrial systems, too much nitrogen can affect species diversity by favoring
14 plant species with high nitrogen requirements over other species which may have evolved
15 specialized capabilities to thrive under nitrogen-limited conditions. Elevated soil nitrogen tends
16 to cause changes in invasive species colonization, biodiversity, and community structure.
17 Animals that depend on specific plants for habitat and food may also be threatened by changes in
18 vegetation resulting from increased nitrogen inputs.

19 Atmospheric deposition of nitrogen is a significant source of total nitrogen to many
20 estuaries in the United States. The amount of nitrogen entering estuaries ultimately attributable to
21 atmospheric deposition is not well defined. On an annual basis, it may be up to 40% of total
22 nitrogen load, depending on the size and location of the watershed. A number of uncertainties
23 may result in a greater relative contribution of atmospheric deposition in some places. In
24 addition, episodic inputs, which may be ecologically significant, may be higher than the annual
25 average. Estuaries in the U.S. that suffer from over-enrichment of nitrogen often experience a
26 condition known as eutrophication. Symptoms of eutrophication include changes in the dominant
27 species of phytoplankton, low levels of oxygen in the water column, fish and shellfish kills,
28 outbreaks of toxic alga, and population changes which can cascade up the food chain. In
29 addition, increased levels of turbidity in the water due to large amounts of algae can kill off
30 submerged aquatic vegetation, which is important habitat for many estuarine fish and shellfish
31 species.

1.4.3 Ammonia

While ammonia (NH_3) is not the focus of this review, the interactions between NH_3 and NO_x and SO_x in the atmosphere and associated deposition processes will be considered. Ammonia, is an additional source of nitrogen to the environment that contributes to the problems caused by excess nitrogen deposition—fertilization and eutrophication. NH_3 can also contribute to increasing soil acidity via nitrification, the microbial mediated transformation of reduced nitrogen to nitrate. The atmospheric lifetime of gas-phase NH_3 is on the order of a day; however, in the presence of sulfuric or nitric acids, ammonia can form aerosols, predominantly ammonium sulfate and ammonium nitrate, with lifetimes on the order of 7-10 days. This longer lifetime gives the potential for long-range transport of nitrogen and sulfur, and contributes to fine particulate matter pollution and regional haze. Ammonium sulfates and nitrates also participate in climate modification effects, directly through radiative cooling by light scattering and indirectly through modifications of cloud cover and precipitation by acting as cloud condensation nuclei. A detailed discuss of these processes is included in the PM review.

2. REVIEW SCHEDULE

EPA's National Center for Environmental Assessment in Research Triangle Park, NC (NCEA-RTP) announced the official initiation of the current periodic review of air quality criteria for NO_x on December 9, 2005 and for SO_x on May 15, 2006. For each of these reviews, the Agency began by announcing in the Federal Register (70 FR 73236 and 71 FR 28023) the formal commencement of the review and a call for information. With these reviews underway, a workshop addressing the separate, joint review of just the secondary standards for these two pollutants was announced in the Federal Register on June 20, 2007 (72 FR 11960). The proposed schedule for this joint review process is shown in Table 2-1; underlined dates indicate the court-ordered schedule.

Table 2-1. Proposed Schedule for Joint NO_x and SO_x Secondary Standard Review¹

Stage of Review	Major Milestone	Draft Target Dates
Planning	Literature Search	Ongoing
	Federal Register Call for Information	December 2005
	Prepare Draft NO ₂ /SO ₂ NAAQS Work Plan	December 2005-August 2007
	Workshop on science/policy issues	July 2007
	CASAC consultation	October 2007
	Prepare final integrated NO ₂ /SO ₂ NAAQS Work Plan	December 2007
Integrated Science Assessment (ISA)	Prepare first draft of ISA	December 2007
	CASAC/public review first draft ISA	March 2008
	Prepare second draft of ISA	July/August 2008
	CASAC/public review second draft ISA	October 2008
	Prepare final ISA	<u>December 12, 2008</u>
Risk/Exposure Assessment (R/EA)	REA methodology released to CASAC and the public	January 2008
	CASAC/public consultation on REA methodology	March 2008
	First draft REA released to CASAC and the public	August 2008
	CASAC/public review of the first draft of the REA	October 2008
	Second draft of REA released to CASAC and the public	March 2009
	CASAC/public review of second draft of REA	May 2009
	Final REA released	July 2009
Policy Assessment/Rulemaking	Publish ANPR	August 2009
	CASAC review/public comment on ANPR	October 2009
	Proposed rulemaking	<u>February 12, 2010</u>
	Final rulemaking	<u>October 19, 2010</u>

¹ Schedule may be modified from time to time, as necessary, to reflect actual project requirements and progress.

3. KEY POLICY-RELEVANT ISSUES

3.1 ISSUES TO BE CONSIDERED IN THE CURRENT REVIEW

In this review of the ecosystem-related public welfare effects of NO_x and SO_x, a series of policy-relevant questions will frame our approach and will be addressed in detail in the science assessment, risk/exposure assessment, and policy assessment sections of the review. For the first time in the NAAQS review process, the secondary standards for NO_x and SO_x will be reviewed together due to the pollutants' combined effects on atmospheric chemistry, deposition processes, and ecosystem-related public welfare effects. As a result, the issue of appropriate indicators becomes central to the review of the standards. This review will evaluate what appropriate indicators are for NO_x and SO_x, if these indicators should be combined into one standard, if separate standards should be issued, or if a suite of standards should be issued. In evaluating environmental responses to these pollutants, the variability present in ecosystems across the nation should also be considered. Issues of ecosystem susceptibility should be addressed, as should the issue of whether individual effects or combined effects are more important to a given ecosystem (i.e. is it NO_x or SO_x alone or the combination that is important).

As noted in the introduction, both EPA and CASAC have acknowledged the importance of NO_x, SO_x, and their associated particulates with respect to acidification effects on ecosystems. This review will focus on the ecosystem-related welfare effects that result from the deposition of these pollutants and transformation products in the gas-phase, rather than on the effects of particulate NO_x and SO_x that remain in the atmosphere. The issues associated with NO_x and SO_x particles; including visibility impairment and climate associated with ambient concentrations of particulates and aerosols are being addressed in the secondary PM NAAQS review, which is also currently underway.

For this secondary NO_x/SO_x review, the primary policy-relevant questions include:

- I. What are the known or anticipated welfare effects influenced by ambient NO_x and SO_x, and for which effects is there sufficient information available to be useful as a basis for considering distinct secondary standard(s)?
- II. What is the nature and magnitude of ecosystem responses to NO_x and SO_x that are understood to have known or anticipated adverse effects and what is the variability associated with those responses (including ecosystem type, climatic

conditions, environmental effects and interactions with other environmental factors and pollutants)?

III. What are the biologically-relevant indices that adequately capture the relationship between ecosystem exposure and response for the known or anticipated welfare effects we are trying to protect?

IV. What are the appropriate air quality indicator(s), averaging time(s), form(s), and level(s) of standards that are requisite to protect those ecosystem responses?

V. To what extent do the current standards provide the requisite protection for the public welfare effects associated with NO_x and SO_x?

a. Should the current secondary standards for NO₂ (as an indicator of NO_x) and SO₂ (as an indicator for SO_x) be retained, revised, or revoked and/or replaced with alternative standard(s) having different indicators to provide the required protection from known or anticipated adverse public welfare effects?

In order to be able to answer these questions, we believe that the relevant scientific issues that need to be addressed in the science, risk/exposure, and policy assessment portions of this review include:

- identifying important chemical species in the atmosphere
- identifying the atmospheric pathways that govern chemical transformation, transport, and deposition of NO_x and SO_x to the environment
- identifying the attributes of ecosystem receptors that govern their susceptibility to deposition of nitrogen and sulfates
- identifying the relevant time scales of ecosystem impacts and matching those time scales to relevant time scales for ambient indicators
- describing the relationships between ambient indicators and biologically relevant indices of effects, including ecosystem services associated with the indicator (but not excluding other non-economic evaluations)
- determining the economic value of ecosystem services as one measure of adversity
- evaluating environmental impacts and sensitivities to varying meteorological scenarios and climate conditions (including if current levels may have a long-term impact due to cumulative loadings, and if this is relevant to a NAAQS review)

- 1 • determining the ambient levels of NO_x and SO_x that provide the requisite protection of
- 2 public welfare

4. SCIENCE ASSESSMENT

4.1 SCOPE AND ORGANIZATION

The scope of the joint NO_x and SO_x secondary NAAQS science assessment is limited to topics that do not duplicate those addressed by the particulate matter (PM) science assessment. The welfare effects associated with NO_x and SO_x criteria pollutants fall into two broad categories: (1) those directly caused by the pollutants or their transformation products while still residing in the ambient air (e.g., visibility impairment and climate interactions) and (2) those driven by the pollutants or their transformation products once deposited onto the environment (e.g. alteration of ecosystems and materials and structures). The welfare effects category addressing NO_x, SO_x and their transformation products while in the ambient air are addressed within the secondary PM NAAQS review, as these processes occur via NO_x and SO_x residing in the particulate or aerosol phase and interact with other chemical components of PM. The focus of the NO_x and SO_x secondary NAAQS joint science assessment will therefore be the latter category of welfare effects that address deposition onto the environment.

The science assessment will be organized into the integrated science assessment (ISA) and its supporting annexes. The ISA will critically evaluate and integrate the scientific information on the exposure and environmental effects associated with atmospheric deposition of NO_x and SO_x to provide a policy-relevant review as discussed in Chapter 3². The annexes, which evaluate and summarize relevant studies, will provide a detailed basis for developing the ISA. The annexes will evaluate thousands of published papers. Key findings will be summarized and integrated by discipline, pollutant impact or assessment endpoint pertinent to decisions on possible revision of the NO₂ and SO₂ secondary NAAQSs. Although emphasis will be placed on the presentation of environmental effects data, other scientific data will also be presented and evaluated in order to provide a better understanding of the nature, sources, distribution, measurement, transformations, and concentrations of NO_x and SO_x in ambient air and in various environmental compartments

The focus of the ISA and its annexes will be literature published since the previous reviews of air quality criteria for NO_x and SO_x. Key findings and conclusions from the previous

² Note that evidence related to human health effects of NO_x and SO_x will be considered in two separate science assessment documents that will be prepared as part of the review of the primary NAAQS.

Air Quality Criteria Documents (AQCDs) will be briefly summarized at the beginning of the ISA. The results of recent studies will be integrated with previous findings. Important older studies will be more specifically discussed if they are open to reinterpretation in light of newer data. Generally, only information that has undergone scientific peer review and that has been published (or accepted for publication) in the open literature will be considered. However, exceptions may be made depending on the importance of the subject information and its relevance to the review of the NO₂ and SO₂ secondary NAAQS, as determined in consultation with CASAC.

Emphasis will be placed on studies conducted at or near NO_x and SO_x concentrations found in ambient air. Other studies may be included if they contain unique data, such as the documentation of a previously unreported effect or of a mechanism for an observed effect; or if they were studies that included both higher and lower concentrations designed to determine exposure-response relationships.

4.2 ASSESSMENT APPROACH

The NCEA-RTP is responsible for preparing the ISA and its annexes for NO_x and SO_x. Expert authors include EPA staff with extensive knowledge in their respective fields and extramural scientists contracted to the EPA.

4.2.1 Literature Search

The NCEA-RTP uses a systematic approach to identify relevant studies for consideration. A Federal Register Notice (FRN) was published on December 9, 2005 for NO_x and on May 15, 2006 for SO_x to announce the initiation of this review and request information from the public. An initial publication base has been established by searching various databases using as key words the terms nitrogen oxides (NO_x), nitrogen dioxide (NO₂), nitric acid (HNO₃), peroxyacetyl nitrate (PAN), total reactive nitrogen, terrestrial ecosystems, aquatic ecosystems, wetlands, etc. This search strategy will periodically be reexamined and modified to enhance identification of pertinent published papers. Additional papers are identified for inclusion in the publication base in several ways. First, EPA staff reviews pre-publication tables of contents for journals in which relevant papers are published. Second, expert chapter authors are charged with independently identifying relevant literature. Finally, additional publications are identified by both the public and CASAC during the external review process. The studies identified will include research

published or accepted for publication by a date determined to be as inclusive as possible given the relevant target dates in the NAAQS review schedule. Some additional studies, published after that date, may also be included if they provide new information that impacts one or more key scientific issues. The combination of these approaches should produce a comprehensive collection of pertinent studies needed to form the basis of the ISA.

4.2.2 Criteria for Study Selection

Emphasis shall be placed on 1) recent U.S. studies; 2) studies that evaluate effects at realistic ambient levels; and 3) studies that consider oxides of nitrogen and sulfur as components of a complex mixture of air pollutants. Studies conducted in other countries that contribute significantly to the knowledge base shall be included in the assessment. In assessing the relative scientific quality of studies reviewed here and to assist in interpreting their findings, the following considerations will be taken into account: (1) to what extent are the aerometric data/exposure metrics of adequate quality and sufficiently representative to serve as credible exposure indicators; (2) were the study populations well-defined and adequately selected so as to allow for meaningful comparisons between study groups; (3) were the ecological assessment endpoints reliable and policy-relevant; (4) were the statistical analyses used appropriate and properly performed and interpreted; (5) were likely important covariates (e.g., potential confounders or effect modifiers) adequately controlled or taken into account in the study design and statistical analyses; and (6) were the reported findings internally consistent, biologically plausible, and coherent in terms of consistency with other known facts.

These guidelines provide benchmarks for evaluating various studies and for focusing on the highest quality studies in assessing the body of environmental effects evidence. Detailed critical analysis of all NO_x and SO_x environmental effects studies, especially in relation to the above considerations, is beyond the scope of the ISA and its annexes. Of most relevance for evaluation of studies is whether they provide useful qualitative or quantitative information on exposure-effect or exposure-response relationships for environmental effects associated with current ambient air concentrations of NO_x and SO_x or deposition levels likely to be encountered in the United States.

4.2.3 Content and Organization of the ISA

The content of the ISA will be organized and developed to address the overarching policy-relevant questions listed in Chapter 3.

The scientific information in the ISA will be drawn from a series of more comprehensive ISA annexes. The annexes will be focused on accomplishing two goals. The first goal will be to identify scientific research that is relevant to informing key policy issues. The second goal will be to produce a base of evidence containing all of the publications relevant to the review of the NO_x and SO_x secondary standards. Sections of the ISA annexes will include the following, not necessarily in this order:

- (1) Atmospheric chemistry of NO_x and SO_x; emissions, transport, and transformations in the atmosphere
- (2) Deposition processes of atmospherically derived NO_x and SO_x, ambient concentrations and the relationship between ambient concentrations and deposition
- (3) Environmental response networks, monitoring networks and models;
- (4) Acidification effects of NO_x and SO_x on terrestrial and aquatic ecosystems
- (5) Non-acidification effects of NO_x on terrestrial and aquatic ecosystems;
- (6) Non-acidification effects of SO_x on terrestrial and aquatic ecosystems;
- (7) Impacts of NO_x/SO_x on ecosystems modified by climate change factors
- (8) Assessing the available information for potentially determining critical loads to ecosystems
- (9) Effects of NO_x and SO_x deposition on acid-driven erosion of man-made materials and structures; and
- (10) Valuation of environmental effects of NO_x and SO_x on ecosystems as one of a number of means of assessing when an effect can be considered adverse

The ISA will convey the reasoning used to select which scientific studies are most policy-relevant. Detailed discussions of studies considered for the ISA, but not deemed most relevant, will be limited to the annexes. The ISA will integrate scientific information from the 10 subject areas identified by the annexes into main topic areas. At this point we anticipate these sections to be Atmospheric Concentration and Deposition, and Ecosystems. However upon further evaluation of the literature Materials and Structures Damage may be a third area of the ISA.

1. ***Atmospheric Concentration and Deposition:*** The ISA will evaluate what we know about the factors that influence deposition, the uncertainties associated with extrapolation from ambient air concentrations to ecosystem exposures to NO_x and SO_x, and spatial patterns of NO_x and SO_x deposition. Specific questions include, but are not limited to:
 - a. How are uncertainties described when extrapolating between stationary monitoring instruments and ecosystem exposure?
 - b. What is the spatial distribution of NO_x and SO_x on a national scale? What are the most effective means of using our observations and predictive models to address spatial, temporal and compositional distribution of nitrogen and sulfur? What additional information (observations, process formulations) would enhance our ability to quantify current exposures and track changes over time?
 - c. How does reduced nitrogen (NH_x) alter the atmospheric chemistry and deposition of NO_x or SO_x?
2. **Integration of *ecological* evidence:** The integration section will begin with an overview of three ecologic concepts that are important to clarify at the onset of the discussion: scale, function and structure (organism to region) and ecosystem services. The integration of ecologic evidence will be organized according to impacts. These impacts are acidification (combined effects of NO_x and SO_x), nitrogen nutrient effects (NO_x) and sulfur nutrient effects (SO_x). With regard to nitrogen nutrient effects, it is important to note that ecosystems are typically sensitive to total reactive nitrogen (Nr) input. Therefore NO_x is evaluated as a component of total nitrogen input from various anthropogenic sources. The following questions will be addressed for each ecological impact:
 - a. Is the biogeochemistry of ecosystems impacted and what are the chemical indicators of impact?
 - b. How are organisms impacted? Which biological species are most sensitive? How is sensitivity defined?
 - c. What ecosystems are most sensitive to NO_x and SO_x pollution?
 - d. How does NO_x and SO_x pollution impact ecosystem services?
 - e. What new information is available on the levels of deposition that can cause harm in ecosystems?
 - f. What are the most appropriate spatial and temporal scales to evaluate impacts on ecosystems?
 - g. What is the relationship between ecological vulnerability to NO_x and SO_x pollution variations in current meteorology or gradients in climate?

4.3 PUBLIC AND SCIENTIFIC REVIEW

4.3.1 Informal Peer Consultation Workshop of the ISA Annexes

A combined peer review/policy kickoff workshop was held in Chapel Hill, NC on July 17-19, 2007. This peer consultation workshop provided a forum for the authors of the ISA and its annexes to receive comments from their scientific peers on the first draft of their findings. Peer reviewers were given a copy of the chapter they were assigned prior to the workshop and were asked to conduct an informal review by providing written comments on its scientific content. Each peer consultant was also asked to participate in the workshop. Reviewers provided critical comments on the adequacy of the coverage of the literature, the presentation of the evidence in both text and figures or tables, the appropriateness of the conclusions from the evidence and any new issues or literature that should be considered. At this workshop, expert reviewers were also asked to participate in a discussion of the integrative aspects of the evidence, to assist the EPA in subsequent preparation of the ISA.

4.3.2 Public Review of External Review Drafts

Since the conclusion of the workshop review process, the authors, contributing reviewers, and NCEA-RTP Project Team have met to resolve how to address comments received and will revise the draft annexes and complete preparation of the First External Review Draft (ERD) NO_x and SO_x ISA. After clearance by the U.S. EPA, the draft document will be released for public comment as announced in a Federal Register Notice. The ERD will be made available for review during a specified time period, usually of 60 to 90 days; written comments are solicited during this time. A similar procedure will be followed for public and CASAC review of a Second External Review Draft that EPA expects to be necessary before completion of the Final NO_x and SO_x ISA.

4.3.3 Review by Clean Air Scientific Advisory Committee

At the time the First External Review Draft is released to the public, that draft document will also be sent to the Clean Air Scientific Advisory Committee (CASAC) of EPA's Science Advisory Board (SAB). CASAC members and consultants (see Appendix A) will review the

1 draft document and discuss their comments in a public meeting announced in the Federal
2 Register. At the meeting, the NCEA-RTP Project Team plans to present an overview of the main
3 features of the document, a summary of key issues raised by public comments received on the
4 document, and the charge to the committee, as well as being prepared to discuss proposed
5 revisions, if indicated. Based on CASAC's past practice, EPA expects that key CASAC advice
6 and recommendations for revision of the document will be summarized by the CASAC Chair in
7 a letter to the EPA Administrator. EPA will take into account any such recommendations, as
8 well as CASAC and public comments at the meeting and any written comments received, in
9 revising the draft NO_x and SO_x ISA. As noted earlier, EPA expects that it will be necessary to
10 prepare a Second External Review Draft for further CASAC review and public comment. After
11 appropriate revision, the final document will be made available on an EPA website and
12 subsequently printed, with its public availability being announced in the Federal Register.

13

5. RISK/EXPOSURE ASSESSMENT

The risk/exposure assessment for the NO_x and SO_x secondary NAAQS review will build upon the scientific information presented in the Integrated Science Assessment. As discussed earlier in this document, the risk/exposure assessment will not focus on other welfare effects that might be associated with secondary pollutants associated with NO_x and SO_x. Depending on the results of the science assessment, air quality indicators for these criteria pollutants that differ from the current indicators may be considered in the risk/exposure assessment.

The risk/exposure assessment will evaluate potential alternative indices, in an attempt to quantify the relationship between ambient concentrations of NO_x and SO_x and potential welfare effects. To create these indices, we will evaluate exposures and impacts in various ecosystems with differing responses related to nitrogen and sulfur inputs. Further, the risk/exposure assessment will use a variety of methods to assess the potential adversity of impacts including the effects of the pollutants on ecosystem goods and services, the degree to which ecosystem functions are impaired, long-term trends in specific ecosystems (where available), and both economic and non-economic valuation of ecosystem services. Our ability to address these issues will enable us to determine the extent to which the current standards provide requisite protection from any known or anticipated adverse effects associated with ambient levels of NO_x and SO_x. The risk/exposure assessment will recognize that the Administrator makes the final decision as to whether an effect is significantly adverse to warrant revising and/or replacing the current standards, or if the body of scientific evidence and assessment tools provide an adequate basis for supporting revised standards.

To evaluate the nature and magnitude of ecosystem responses associated with adverse effects, the risk/exposure assessment will examine various ways to quantify the relationship between air quality indicators, deposition of biologically-accessible forms of nitrogen and sulfur, biologically-relevant indices relating deposition, exposure and effects on sensitive receptors and related impacts to ecosystem service(s). The risk/exposure assessment will evaluate the overall load to the system for nitrogen and sulfur as well as the variability in ecosystem responses to these pollutants and determining the exposure metric(s) that incorporates the temporal considerations (i.e. biologically-relevant timeframes), pathways, and biologically-relevant indices in order to maintain the functioning of these ecosystems. In addition, the risk/exposure assessment will attempt to address how sensitive ecosystems and their services affected by

1 nitrogen and sulfur deposition may respond to variations in climatic elements such as
2 temperature and rainfall.

3 The scope of the risk/exposure assessment will depend in part on the answers to the
4 following questions:

- 5 • What are the appropriate geographic scale(s) and/or time frame(s) for the risk
6 assessment? Information that will be considered in addressing this question
7 includes: national-scale mapping, case studies of sensitive ecosystems, identification
8 of representative ecosystem types, a weight-of-evidence approach incorporating both
9 qualitative and quantitative information on risk, and perhaps some combination of
10 the above.
- 11 • How can regional variation of effects be taken into account? How should the
12 risk/exposure assessment address acidification and fertilization effects in different
13 areas and ecosystem types at varying ambient concentrations of nitrogen and sulfur
14 compounds? Can differences in regional sensitivity to deposition be completely
15 characterized by underlying physical or biological attributes?
- 16 • To what degree can assumptions be made about linkages between pollutants in
17 ambient air, deposition, and measurable ecosystem effects? What are the most
18 useful metrics of both ambient pollution and the resulting effect?
- 19 • To what degree should the risk/exposure assessment take the potential for recovery
20 into account in selecting data for qualitative and quantitative assessments?
- 21 • How can uncertainties be minimized and appropriately characterized?

23 **5.1 ASSESSMENT APPROACH**

24 There are a variety of ways to address the risk/exposure assessment for the review of the
25 secondary NAAQS for NO_x and SO_x. One of the purposes of this document is to solicit ideas
26 and evaluate the most appropriate way to approach this review. In order to assess risks to public
27 welfare due to ambient nitrogen and sulfur deposition, the relationships (and associated
28 uncertainties) between ambient concentrations, deposition, environmental effects, environmental
29 receptors, and ecosystem services and ecosystem valuation should be established. Some of our
30 preliminary ideas for the risk/exposure assessment are outlined below. A detailed description of
31 the tools mentioned here and others that are currently available are described in detail in
32 Appendix B.

1 Understanding the atmospheric chemistry and deposition of nitrogen and sulfur is central
2 to this review. Most likely, we will use the Community Multi-Scale Air Quality (CMAQ) model
3 (and derivative models such as the response-surface model) described in Appendix B to estimate
4 ambient concentrations and deposition areas. Any assessment of atmospheric chemistry and
5 deposition is highly-dependent on the variability associated with baseline inputs. Consideration
6 will be given to analyzing five consecutive years of CMAQ outputs (2001-2005) delivered as
7 part of EPA's Interagency Agreement with CDC to improve understanding of baseline
8 deposition and the relationships between nitrogen and sulfur concentration and deposition in
9 terms of magnitude and spatial correlations.

10 Initially we plan to conduct regression and related statistical analyses to generate
11 statistical relationships between grid-level ambient concentrations of NO_x and SO₂ and grid-level
12 or appropriately spatially averaged (e.g. at the watershed level) deposition rates of sulfur and
13 nitrogen. The regression analysis will use the results from the 2001-2005 CMAQ simulations
14 conducted for the joint EPA/CDC project. The analysis will use linear and non-linear regression
15 modeling, as well as spatial regression techniques, to develop reduced form equations. To the
16 extent that data are available, the statistical modeling may also incorporate relevant
17 meteorological and physical attributes, e.g. precipitation, relative humidity, wind speed and
18 direction. These statistical relationships are not expected to provide predictive models. Rather,
19 the goal is to develop appropriate estimates of the relationship between deposition metrics and
20 ambient levels of NO_x and SO_x.

21 Another central issue in this review is determining which ecosystems are more or less
22 sensitive to nitrogen and sulfur deposition. It may be useful to normalize all ecosystems (where
23 data is available) based on a small subset of determinative underlying characteristics. These
24 characteristics may include (but are not limited to): (1) potential nitrogen and sulfur retention
25 rates, (2) potential nitrogen and sulfur uptake rates, which might include vegetative uptake,
26 potential denitrification, and potential mobilization of nitrogen and sulfur, (3) potential residence
27 time based on local hydrology (precipitation rates, conductivity) and geology (bedrock type,
28 pervious surfaces, soil type and characteristics), and (4) total supply of nitrogen and sulfur
29 including atmospheric deposition, and other non-atmospheric sources (such as fertilization,
30 sewer leaks, point sources, etc.). Other ecosystem-specific characteristics which may help assess
31 sensitivities include threatened and endangered species data where available, land-use type
32 (including Class I Wilderness areas), and baseline nitrogen and sulfur loading estimates. Where

case-study or ecosystem-specific data are available, a subset of maps for the case-study region may be created. Complementary to these efforts, we may use a statistical cluster analysis to group ecosystem units into similar sets. By clustering ecosystems together, we may reduce the number of locations that need to be modeled to adequately characterize the variability in ecosystem response to changes in nitrogen and sulfur deposition.

These types of analyses may aid in determining if case studies might be appropriate for looking at nitrogen and sulfur effects on various ecosystems and geographic regions as a means of extrapolating these impacts to the entire country. For those areas where data are available, watershed models (e.g., MAGIC, PnET-BGC, and DayCent-Chem) may be useful for evaluating the emission-deposition-ecosystem response linkage. (These models are described in more detail in Appendix B.) As the integrated science assessment and risk/exposure assessment progress, we anticipate identifying additional ways to evaluate biologically-relevant indicators and ecosystem valuation techniques that will ultimately help to direct and focus the final Risk Assessment.

5.2 TOOLS FOR RISK ASSESSMENT

In order to develop a risk analysis for a meaningful nationwide or ecosystem-specific metric, some of the available analytical tools for conducting a risk assessment are described/summarized in Appendix B. The more detailed model evaluation discussion included in the annexes and ISA will aid in selecting the appropriate modeling tools. With these tools, the risk/exposure assessment should be able to determine the appropriate endpoint(s)/indicator(s), geographic level/scale of protection, national or case-study modeling, and which ecosystem services are important.

5.3 KNOWLEDGE GAPS

As discussed above and in detail in Appendix B, the risk/exposure assessment will use numerical models of atmospheric emissions, transformation, fate, transport, and deposition to follow emissions from a source to a receptor, either aquatic or terrestrial. We plan to overlay available data sets related to deposition and adverse impacts with GIS-based approaches to aid in the evaluation of their correlations. We also have published and evaluated regional or area-specific models that will aid in evaluating ecosystem effects in specific watersheds (i.e. for the Chesapeake Bay or the Adirondack mountains). Additionally, we are beginning to quantify

ecosystem services and the benefits they provide. However, there are knowledge gaps in the linkages between these areas.

Some of the main questions the risk/exposure assessment will try to answer are, “If we can identify appropriate biologically relevant indices, can we establish a link between deposition of NO_x and SO_x and ecosystem response? Additionally, what ecosystem services are associated with changes in emissions?” Answering these questions, raises other issues regarding the links between atmospheric deposition and an ecosystem response, and ecosystem services and emissions changes. Some of these questions are listed below.

Atmospheric Deposition, Fate, and Transport Modeling

- What are typical uncertainties for observed and regional model estimates of wet and dry NO_x, NH_x, and SO_x deposition rates?

- Can a deposition model be used to estimate deposition to a land-use specific surface, such as mountainous terrain, coniferous/deciduous forests, etc?

- If not, are there models that can take this output and estimate deposition to specific ecosystems?

Ecological Response Modeling

- Do we have sufficient data to develop dose-response functions linking changes in nitrogen and sulfur deposition to changes in ecosystem functions?

- Can we adequately characterize the uncertainties in these dose-response functions?

- Can ecosystem-specific models be used to represent similar ecosystems nationwide?

What are the uncertainties associated with the transferability between a specific watershed model to multiple watersheds?

Ecosystem Services

- Can the current known ecological responses be translated into an ecosystem service?

- Can a process-based model serve to link multiple ecosystem services in order to assess the ‘true’ impact of air quality on the ‘bundles of services’ provided by that ecosystem?

- Can we value the ecosystem impacts and benefits associated with a given indicator?

- How can we show that reducing emissions of NO_x/SO_x results in a specific ecosystem benefit?

1 *Valuation*

- 2 • Is there an adequate database of economic values for changes in ecosystems or
3 ecosystem services?
4 • Are there other methods for developing non-monetized value estimates?
5 • How are valuation methodologies to be combined with non-valuation means of
6 assessing adversity of effect?
7

8 **5.3 PUBLIC AND SCIENTIFIC REVIEW**

9 Drafts of the risk/exposure assessment will be reviewed by CASAC of EPA's Science
10 Advisory Board (SAB). CASAC members and consultants (Appendix A) will review the draft
11 document and discuss their comments in a public meeting announced in the Federal Register.
12 Based on CASAC's past practice, EPA expects that key CASAC advice and recommendations
13 for revision of the document will be summarized by the CASAC Chair in a letter to the EPA
14 Administrator. In revising the draft risk/exposure assessment for NO_x and SO_x, EPA will take
15 into account any such recommendations. EPA will also consider comments received, from
16 CASAC or from the public, at the meeting itself and any written comments received. EPA
17 anticipates preparing a second draft of the risk/exposure assessment for CASAC review and
18 public comment. After appropriate revision, the final document will be made available on an
19 EPA website and subsequently printed, with its public availability being announced in the
20 Federal Register.

6. POLICY ASSESSMENT

The Policy Assessment will synthesize the information from the ISA and the risk/exposure assessment to evaluate if the current standards provide requisite protection for the welfare effects associated with NO_x and SO_x, and if the current standards should be retained, revised, or revoked. Based on our current understanding of the nature and magnitude of ecosystem responses to these pollutants, the uncertainties associated with them, and the relationship between ambient concentrations and these ecosystem responses, the policy assessment will discuss the appropriate form(s) of standard(s) that will address the ecosystem responses we are trying to protect.

The Administrator will make the final decision on the requisite standard(s) and this will reflect public policy judgments. A final decision must draw upon scientific evidence and analyses about effects on public welfare and the nature and magnitude of ecosystem responses that are understood to have adverse effects, as well as judgments about how to deal with the range of uncertainties that are inherent in this information. This evaluation should recognize that the effects of NO_x/SO_x compounds on aquatic and terrestrial ecosystems are diverse and occur over time as a result of deposition, include acidification and fertilization effects, are regionally specific, and occur through the atmospheric reactions, transport, and deposition of NO_x/SO_x compounds emitted from varied and ubiquitous sources. The links in the fate, transport, and deposition apply to both acidification and fertilization (which is understood to capture both eutrophication in aquatic systems and fertilization in terrestrial systems).

Other considerations in developing an indicator include using total nitrogen as a biologically-relevant index for fertilization. Total nitrogen deposited (and transported) should be viewed in context of baseline nitrogen abundance. Ecosystem impacts may differ due to form of nitrogen deposited (e.g., oxidized vs reduced nitrogen), and the various forms of nitrogen deposited may be important in determining the form of the standard. Nitrogen deposition impacts also depend on depth of soils, species composition and competition, and soil composition. In aquatic ecosystems, the range of biologically-relevant indices for nitrogen sensitivity may include dissolved oxygen concentrations, and the ratio of algae to diatoms. Acid neutralizing capacity (ANC) may be an appropriate biologically-relevant index for acidification depending on the time scale being considered. Acidification measured by sulfate and nitrate concentrations can change relatively rapidly (over several years), while ANC can take decades to change and reach

1 a new equilibrium depending on releases of sulfate due to current emissions/deposition or release
2 of soil-stored sulfate. The short-term effects from acid pulses may point to the need for a
3 seasonal form of the standards. Nitrogen impacts through fertilization also have a seasonal
4 component that may display a different pattern than that for acidification. The need for seasonal
5 standards may differ by the water type and land use in different areas. Ecosystem recovery time
6 may also play an important role in determining the appropriate form and averaging time of
7 standards.

8 The policy determination of what is an adverse impact may vary, depending on the
9 method chosen to assess adversity. The NAAQS provisions in the Act require the Administrator
10 to establish secondary standards that are requisite to protect public welfare from any known or
11 anticipated adverse effects associated with the presence of the pollutant in the ambient air. In so
12 doing, the Administrator seeks to establish standards that are neither more nor less stringent than
13 necessary for this purpose. The provisions do not require that secondary standards be set to
14 eliminate all welfare effects, but rather at a level that protects public welfare from those effects
15 that are judged to be adverse. In the evaluation of ecosystem services and valuation as one
16 potential measure of adversity, the assessment should consider the ecosystem benefits that are
17 projected to occur as a result of reductions in the current standards.

18 In the policy assessment, a series of questions will frame our approach to reaching
19 conclusions based on available evidence and information, as to whether consideration should be
20 given to retaining or revising the current secondary NO₂ and SO₂ NAAQS. The review of the
21 adequacy of the current standard begins by considering whether the currently available body of
22 evidence assessed in the NO_x/SO_x ISA suggests that revision of any of the basic elements of the
23 NAAQS is appropriate. This evaluation is completed for each category of NO_x/SO_x-related
24 welfare effects identified in the NO_x/SO_x ISA associated with the deposition of NO_x/SO_x
25 compounds originating from ambient air. The ability of the risk/exposure assessment to create a
26 link between atmospheric concentrations and associated deposition with ecosystem responses
27 related to biologically-relevant indices will play a significant role in informing the policy
28 determination for this review.

29 The review of the adequacy of the current standard for each effects category involves
30 addressing questions such as:

- To what extent does the available information demonstrate or suggest that NO_x/SO_x-related effects are occurring at current ambient conditions or at levels that would meet the current standards?
- To what extent does the available information inform judgments as to whether any observed or anticipated effects are adverse to public welfare?
- To what extent are the current secondary standards likely to be effective in achieving protection against any identified adverse effects?

To the extent that the evidence suggests that revision of the current secondary NO₂/SO₂ NAAQS is appropriate, ranges of standards will be identified (including different or alternate indicators, terms of exposure indices, averaging times, levels, and forms) that reflect a range of alternative policy judgments as to the degree of protection that is requisite to protect public welfare from known or anticipated adverse effects. To account for variability in ecosystem responses across the nation, ecosystem characteristics may be an important consideration in evaluating the form(s) of the standard(s). The form(s) of the standard(s) may be based on a complex formula that incorporates ecosystem characteristics, atmospheric transformations, climatic conditions, environmental effects and other interactions. In so doing, the following questions should be addressed:

- Does the available information provide support for considering different NO_x/SO_x chemical indicators or exposure indices?
- Does the available information provide support for considering some joint standard(s) or are separate standards appropriate?
- What range of levels and forms of alternative standards are supported by the information, and what are the uncertainties and limitations in that information?
- To what extent do specific levels and forms of alternative standards reduce adverse impacts attributable to NO_x/SO_x, and what are the uncertainties in the estimated reductions?

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8. APPENDIX A

U.S. EPA SCIENCE ADVISORY BOARD CLEAN AIR SCIENTIFIC ADVISORY COMMITTEE MEMBERS

FISCAL YEAR 2007

The Clean Air Scientific Advisory Committee (CASAC) has a statutorily mandated responsibility to review and offer scientific and technical advice to the Administrator on the air quality criteria and regulatory documents that form the basis for the national ambient air quality standards (NAAQS), which currently include standards for lead (Pb), particulate matter (PM), ozone (O₃), carbon monoxide (CO), nitrogen dioxide (NO₂) and sulfur dioxide (SO₂). To perform such reviews, in each case the Committee forms a review panel consisting of CASAC members augmented by selected consultants with expertise in scientific or technical areas pertinent to the given pollutant or pollutant class under review.

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15 University of North Carolina at Chapel Hill
16 Chapel Hill, NC 27599
17

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23

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35

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32 Washington, DC 20460

33
34 Ms. Zisa Lubarov-Walton

35 Management Assistant

36 Science Advisory Board (1400A)

37 US Environmental Protection Agency

38 Washington, DC 20460

9. APPENDIX B

B.1 MODELS

Community Multi-Scale Air Quality (CMAQ) Model

The CMAQ model is a three-dimensional grid-based Eulerian air quality model designed to estimate the formation and fate of ozone and other oxidants and their precursors; primary PM and secondary particulate matter precursors and atmospheric concentrations; toxics; and deposition of chemical species over scales ranging from continental to regional and urban to neighborhood (EPA, 1999; Byun and Schere, 2006; Dennis et al., 1996). The CMAQ model was peer-reviewed in 2003 for EPA as reported in “Peer Review of CMAQ Model” (Amar et al., 2004).

CMAQ is most often configured to output spatial fields of gridded concentrations and deposition on an hourly basis for the entire modeling domain. Pollutant concentrations are output for each vertical layer included the model simulation. Additional information on the horizontal and vertical configuration of the model is provided below. The current version of CMAQ (v4.6) is capable of one atmosphere modeling in which NO_x, SO_x, ozone, particulates, mercury, and selected toxic pollutants are included in a model simulation. A list of the key nitrogen and sulfur containing deposition species output by CMAQ is provided in Table B.1. This table also identifies the deposition species derived from CMAQ’s nitrogen and sulfur deposition outputs.

The standard hourly CMAQ model predictions are post-processed to create gridded fields of daily average, monthly average, and annual average concentrations for layer 1, which is the layer nearest to the ground. The hourly deposition outputs are post-processed to produce gridded fields of daily, monthly, and annual total wet, dry and wet plus dry deposition.

Table B.1. Key Predicted and Derived Nitrogen and Sulfur Deposition Species from CMAQ¹

<i>Predicted Deposition Species</i>	<i>Derived Deposition Species</i>
Nitrogen Oxide (NO)	Oxidized Nitrogen
Nitrogen Dioxide (NO ₂)	Reduced Nitrogen

Nitric Acid (HNO ₃)	Total Nitrogen
Dinitrogen Pentoxide (N ₂ O ₅)	Total Sulfur
Peroxyacetyl Nitrate (PAN)	
Ammonia (NH ₃)	
Sulfur Dioxide (SO ₂)	
Particulate Sulfate (SO ₄)	
Particulate Nitrate (NO ₃)	
Particulate Ammonium (NH ₄)	

¹Model predictions include both wet and dry deposition in units of kg/ha.

2002-Based Platform CMAQ Modeling

EPA/AQAD is developing an air quality modeling platform that includes meteorology and emissions³ for 2002 along with projected emissions for 2009, 2014, 2020, and 2030. The future year projections reflect the combined effects of emissions growth and reductions associated with “national rules” up through the CAIR/CAMR/CAVR. CMAQ will be run for each of these five years for a 36 km nationwide domain and for separate 12 km modeling domains covering the eastern and western U.S. The area covered by each of these domains is shown in Figure B.1. The domain specifications are provided in Table B.2. All three modeling domains contain 14 vertical layers with a top at about 16,200 meters, or 100 mb. Note that the horizontal and vertical resolution selected for these simulations are not limited by the model. Rather, they were selected to balance (1) the desire to have annual nationwide simulations with fine scale resolution versus (2) the computational requirements of the simulations which increase with horizontal and vertical resolution⁴.

The CMAQ meteorological input files were derived from simulations of the Pennsylvania State University / National Center for Atmospheric Research Mesoscale Model (Grell, Dudhia, and Stauffer, 1994). This model, commonly referred to as MM5, is a limited-area, nonhydrostatic, terrain-following system that solves for the full set of physical and thermodynamic equations which govern atmospheric motions. The outputs from MM5 were

3 Emissions inputs for to CMAQ include anthropogenic and biogenic emissions from areas of the U.S., Canada, and Mexico that are within the modeling domain. Anthropogenic emissions for the U.S. were derived from the 2002 National Emissions Inventory.

4 For various special studies, CMAQ has been run at 4 km and 1 km horizontal resolution for urban-scale domains and with a vertical resolution of over 30 layers.

processed to create model-ready inputs for CMAQ using the Meteorology-Chemistry Interface Processor (MCIP) version 3.2: horizontal wind components (i.e., speed and direction), temperature, moisture, vertical diffusion rates, and rainfall rates for each grid cell in each vertical layer (EPA, 1999).

The lateral boundary and initial species concentrations for the CMAQ 36 km domain were obtained from a three-dimensional global atmospheric chemistry model, the GEOS-CHEM model (Yantosca, 2004). The global GEOS-CHEM model simulates atmospheric chemical and physical processes driven by assimilated meteorological observations from the NASA's Goddard Earth Observing System (GEOS). This model was run for 2002 with a grid resolution of 2 degree x 2.5 degree (latitude-longitude) and 20 vertical layers. The predictions were used to provide one-way dynamic boundary conditions at 3-hour intervals and the initial concentration field for the CMAQ simulations. The outputs from the 36 km domain are utilized to provide the initial and boundary condition concentrations for each of the two 12 km domains.

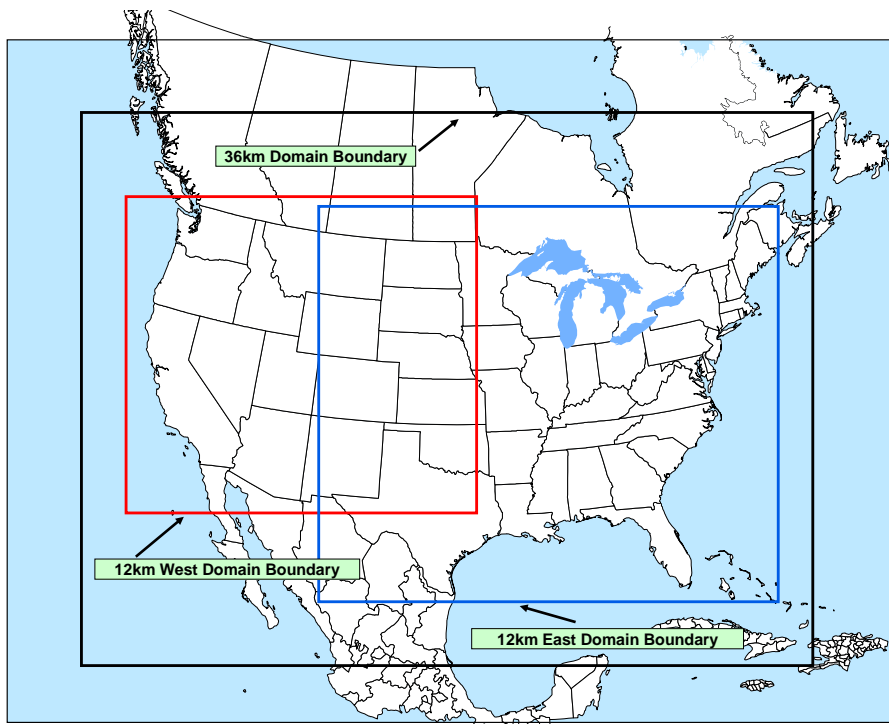


Figure B.1. Map of the CMAQ 36 km and 12 km Modeling Domains.

Table B.2. Geographic Specifications of Modeling Domains.

36 km Domain	12 km Eastern Domain	12 km Western Domain
September 2007	B-3	DRAFT-DO NOT QUOTE OR CITE

(148 x 112 Grid Cells)			(279 x 240 Grid Cells)			(213 x 192 Grid Cells)		
	Lon	lat		lon	lat		lon	lat
SW	-121.77	18.17	SW	-106.79	24.99	SW	-121.65	28.29
NE	-58.54	52.41	NE	-65.32	47.63	NE	-94.94	51.91

Response Surface Models (RSM)

Air quality models can be a powerful regulatory tool for comparing the efficacy of various emissions control strategies, regulatory impacts and policy decisions. However, due to the often large computational costs and the complication of the required emission inputs and processing, using individual simulations from photochemical air quality models to meet policy deadlines can be inefficient. A promising tool for addressing this issue, Response Surface Modeling (RSM), has been developed by using advanced spatial statistical techniques to characterize the relationship between model outputs and input parameters in a highly economical manner. The RSM aggregates numerous pre-specified individual air quality modeling simulations into a multi-dimensional air quality "response surface." Simply, this metamodeling technique is a "model of the model" and can be shown to reproduce the results from an individual modeling simulation with little bias or error. The RSM provides a wide breadth of model outputs that can be used to rapidly assess air quality impacts of different control measures and assist in the development of multi-pollutant control strategies. Specifically, the RSM can be used in a variety of ways: (1) strategy design and assessment (e.g. comparison of urban vs. regional controls; comparison across sectors; comparison across pollutants); (2) optimization (develop optimal combinations of controls to attain standards at minimum cost); (3) model sensitivity (systematically evaluate the relative sensitivity of modeled air quality levels to changes in emissions inputs).

Response surface models have been successfully applied in the context of the ozone and particulate matter regulatory impact analyses. In those contexts, the RSMs were used to explore the sensitivity of ambient ozone and PM concentrations to changes in precursor emissions, for the purpose of designing strategies to reduce ambient concentrations. Documentation of these applications can be found in the regulatory impact analyses (U.S. EPA, 2006a, 2007).

Response surface models for nitrogen and sulfur deposition are in development and should be available to inform the Risk/Exposure Assessment. These models will allow us to explore the sensitivity of nitrogen and sulfur deposition for a number of different deposition

species (reduced and oxidized nitrogen, wet and dry sulfur) to emissions of NO_x, SO₂, and NH₃. The RSMs will also allow us to explore the chained relationships between emissions of NO_x, SO₂, and NH₃ and ambient concentrations of NO_x, NH_x, and SO_x, and the resulting changes in deposition. This will help to identify correlations or differences in the spatial patterns of the response of ambient concentrations versus the response of deposition.

Air Quest

AirQuest is a data management system. It stores air data in a single location and delivers it to various tools and applications. The data in AirQuest can be accessed via SAS, Excel, MS Access, Google Earth, ArcGIS, and a customized web-based interactive map.

AIRQuest was designed to serve as a single source for integrated, high quality, readily available air quality data, summaries, and statistics from regulatory monitoring networks. In addition, AIRQuest represents an opportunity to leverage data processing and storage resources, since a fully functional AIRQuest will remove the need for analysts to store large amounts of data on their desktops. In addition, by carefully defining and coming to agreement on business rules that define “the right way” to calculate air quality metrics, AIRQuest can ensure that every analyst has a consistent approach and consistent answers.

Total Risk Integrated Methodology (TRIM)

The TRIM design includes three individual modules which have been the subject of peer review and publication. Two of these modules are applicable to ecological risk assessments. The *Environmental Fate, Transport, and Ecological Exposure* module, TRIM.FaTE, accounts for movement of a chemical through a comprehensive system of discrete compartments (e.g., media and biota) that represent possible locations of the chemical in the physical and biological environments of the modeled ecosystem and provides an inventory, over time, of a chemical throughout the entire system.

TRIM.FaTE is a spatially explicit, compartmental mass balance model that describes the movement and transformation of pollutants over time, through a user-defined, bounded system that includes both biotic and abiotic compartments. Outputs include pollutant concentrations in multiple environmental media and biota, and also biota and pollutant intakes (e.g., mg/kg-day),

all of which provide exposure estimates for ecological receptors (i.e., plants and animals). Significant features of TRIM.FaTE include: (1) a fully coupled multimedia model; (2) user flexibility in defining scenarios, in terms of the links among compartments, and number and types of compartments, as appropriate for the application spatial and temporal scale; (3) transparent, user-accessible algorithm and input library that allows the user to review and modify how environmental transfer and transformation processes are modeled; (4) a full accounting of all of the pollutant as it moves among environmental compartments during simulation; (5) an embedded procedure to characterize uncertainty and variability; and (6) the capability to provide exposure estimates for ecological receptors.

In the *Risk Characterization* module, TRIM.Risk, estimates of human exposures or doses are characterized with regard to potential risk using the corresponding exposure- or dose-response relationships. The TRIM.Risk module is also designed to characterize ecological risks from multimedia exposures. The output from TRIM.Risk is intended to include documentation of the input data, assumptions in the analysis, and measures of uncertainty, as well as the results of risk calculations and exposure analysis.

The uncertainty and variability features (e.g., sensitivity & Monte Carlo analysis) augment TRIM's capability for performing iterative analyses. For example, the user may perform assessments varying from simple deterministic screening analyses using conservative default parameters to refined and complex risk assessments where the impacts of parameter uncertainty and variability are assessed for critical parameters.

Regional Vulnerability Assessment (ReVA)

The Regional Vulnerability Assessment (ReVA) program conducts research on innovative approaches to the evaluation and interpretation of large and complex datasets and models to assess the current conditions and likely outcomes of environmental decisions, including alternative futures. ReVA works with select client groups to develop research and demonstration pilots on how current data and appropriate models can be combined and interpreted across a geographic region to set management and ecosystem protection priorities and to proactively assess the outcomes of decisions that may impact multiple outcomes or involve tradeoffs in a transparent, defensible fashion.

To date, ReVA has completed the development of an environmental toolkit for EPA Region 3 and is actively engaged in developing tools for decision making that support urban and interstate issues (SEQL), regional multipollutant issues that affect human health and ecosystems (Region 4), and multistate partnerships that affect the health of international environmental resources (Region 5/ Great Lakes).

B.1.2 Case Study Modeling

Integrated assessment case studies provide results for numerous environmental parameters, including emissions, atmospheric concentrations of pollutants, pollutant deposition to land and water surfaces, and changes in ecological parameters such as surface water chemistry, forest soil chemistry, or fish populations. In some cases, it may also be possible to link changes in ecological parameters to changes in ecosystem services, and ultimately to the economic value of the changes in ecosystem services (which is one potential measure of adversity of effect). By comparing the environmental impacts of existing programs with alternative policy scenarios, integrated assessments of this type provide estimates of the incremental ecological effects of potential policy changes. The process for assessing the ecological outcomes of different potential standards frequently focuses on relationships between pollutant emissions, pollutant deposition (the means of exposure for many aquatic and terrestrial ecosystems) and ecosystem response. Currently, analytical tools exist to analyze the emission-deposition-ecosystem response linkage with particular focus on two ecological parameters – surface water chemistry and forest soil chemistry – in three case study areas:

- Northern New England lakes;
- Adirondack Mountain region lakes;
- Southern Blue Ridge region streams.

The assessment approach involves several steps and employs various analytical and modeling tools that, once integrated, provide a means to assess ecological impacts. The National Emissions Inventory is used to develop emissions inventories for area, mobile, and point sources other than electricity generating utilities (EGU). The Integrated Planning Model (IPM) is used to develop emissions inventories for EGUs. The inventories specify profiles of emissions by pollutant, place, time, and source type for each policy scenario. Atmospheric models (e.g., CMAQ) employ the emissions inventory inputs to project changes in atmospheric concentrations

1 and deposition of pollutants to land and water surfaces. Ecological impacts, such as lake and
2 stream water chemistry, or changes in forest soils, are assessed by using watershed models (e.g.,
3 MAGIC, PnET-BGC). These models employ atmospheric modeling output (deposition changes)
4 to project changes in the surface water chemistry of lakes and streams, or changes in forest soil
5 chemistry, in regions that are sensitive to atmospheric deposition of sulfur and nitrogen species.
6 Ecological impacts of policy scenario implementation are then assessed using algorithms based
7 on observed relationships between ecosystem chemical parameters (e.g., acid neutralizing
8 capacity, soil base saturation) and biological indicators of ecosystem health (e.g., fish species
9 richness).

10 Briefly, MAGIC is a lumped-parameter model of intermediate complexity, developed to
11 predict the long-term effects of acidic deposition on surface water chemistry (Cosby et al.
12 1985a,b,c, 2001). The model simulates soil solution and surface water chemistry to predict
13 average concentrations of the major ions. MAGIC calculates for each time step (in this case year)
14 the concentrations of major ions under the assumption of simultaneous reactions involving
15 sulphate adsorption, cation exchange, dissolution-precipitation-speciation of aluminium and
16 dissolution-speciation of inorganic and organic carbon. MAGIC accounts for the mass balance of
17 major ions in the soil by accounting for the fluxes from atmospheric inputs, chemical weathering,
18 net uptake in biomass and loss to runoff.

19 At the heart of MAGIC is the size of the pool of exchangeable base cations in the soil. As
20 the fluxes to and from this pool change over time owing to changes in atmospheric deposition,
21 the chemical equilibria between soil and soil solution shift to give changes in surface water
22 chemistry. The degree and rate of change of surface water acidity thus depend both on flux
23 factors and the inherent characteristics of the affected soils. Data inputs required for calibration
24 of MAGIC comprise lake and catchment characteristics, soil chemical and physical
25 characteristics, input and output fluxes for water and major ions, and net uptake of base cations
26 by vegetation.

27 PnET-BGC is a comprehensive forest-soil-water model developed by linking a monthly
28 carbon, nitrogen, and water balance model PnET (Aber et al., 1997) with a soil model BGC to
29 allow for comprehensive simulations of element cycling within forest and the interconnected
30 aquatic ecosystems (Gbondo-Tugbawa et al., 2001). The model is able to simulate both abiotic
31 processes and biotic processes. Especially the representation of biomass accumulation and the
32 associated element cycling enable the evaluation of land disturbance and climatic events on soil
September 2007

and water chemistry (Gbondo-Tugbawa et al., 2001). The model uses relatively simple formulations and requires a moderate number of inputs to quantify the acid-base status of soil and surface waters under various levels of atmospheric deposition; therefore it can serve both as a research and a management tool. Its simplicity also makes it a good candidate for regional applications.

B.1.3 Ecosystem Services: Impact and Valuation

The final step of an integrated assessment of risks to the public welfare involves determining what level of ecosystem response translates into an effect that could reasonably be considered adverse to the public welfare. There are a number of ways to do this, including

- direct measures of quantities that are of known value to the public, e.g. numbers of endangered species,
- direct economic valuation of ecosystem functions, including use and nonuse values (values that do not require an individual's direct use of an ecosystem – for example, the value of preserving an endangered species habitat, even though that individual will ever see that species in the wild),
- translation of ecosystem attributes into measures of ecosystem services which can then be valued using economic valuation methods, and
- direct non-economic valuation of ecosystem functions based on enumeration of preferences using non-monetized indices of preferences.

Similar to other elements of the Risk/Exposure Assessment, we are seeking input on the range of valuation methods that should be considered in helping to inform the discussion of adversity of ecosystem impacts.

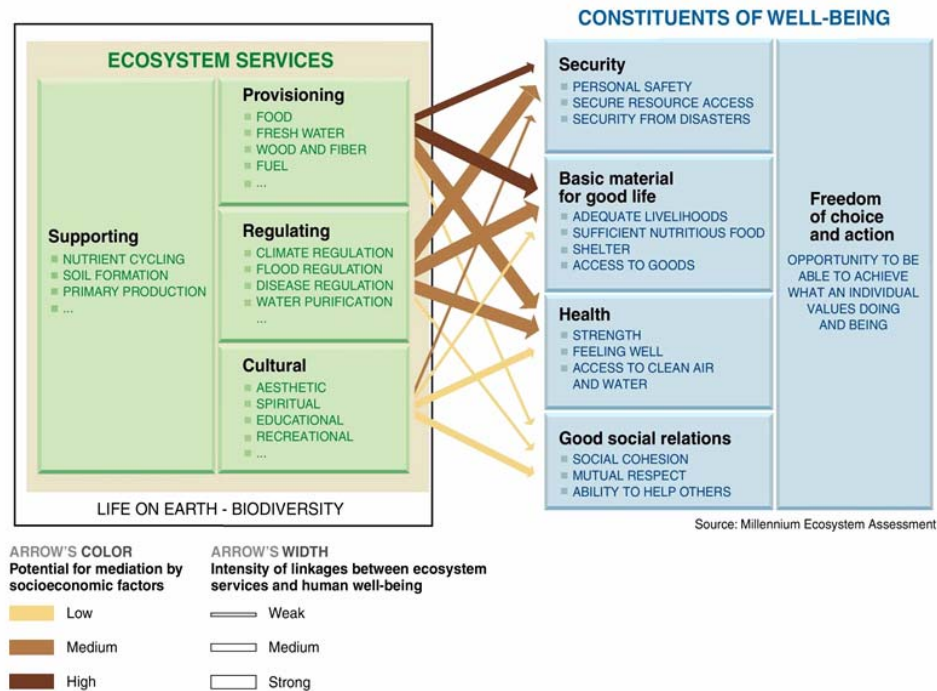
B.1.3.1 Ecosystem services overview and why their consideration is needed

One way to assess adverse effects on welfare is through valuation of ecosystem services. The adverse effect would be the loss or reduction of those services through the effects of NO_x and SO_x on the underlying ecological processes and functions that constitute the service. The

1 EPA defines ecosystem services as the outputs of healthy, intact ecosystems and the underlying
2 ecosystem processes and functions that contribute to human well-being (USEPA 2006b). As
3 articulated by the Millenium Ecosystem Assessment (2005) from the United Nations, these
4 include provisioning services (e.g., clean water, food, wood, fiber, fuel), regulating services (e.g.,
5 water purification, climate regulation, etc.), supporting services (e.g. nutrient cycling, soil
6 formation), and cultural services (e.g. recreation, spiritual) (Figure B.2). Regulating services are
7 of key importance to the EPA because they directly impact air and water quality, and they have
8 strong links to human heath and well-being. As such, valuation of ecosystem services may be
9 one means of assessing whether an effect is adverse. Valuation provides a means for evaluating
10 different services. The valuation may be monetized or non-monetized. Non-monetized measures
11 of valuation include such things as quality of critical habitat, biodiversity, species composition,
12 control/limiting invasive species and pest outbreaks. Determining the exposure response
13 relationships of NO_x and SO_x on the ecosystem service and the underlying ecological process
14 and function will provide a broader focus in determination of adverse impacts.

15 Figure B.2. Ecosystem services as defined by the Millennium Assessment and their linkage to
16 human well being (From Millennium Assessment, 2005).

Focus: Consequences of Ecosystem Change for Human Well-being



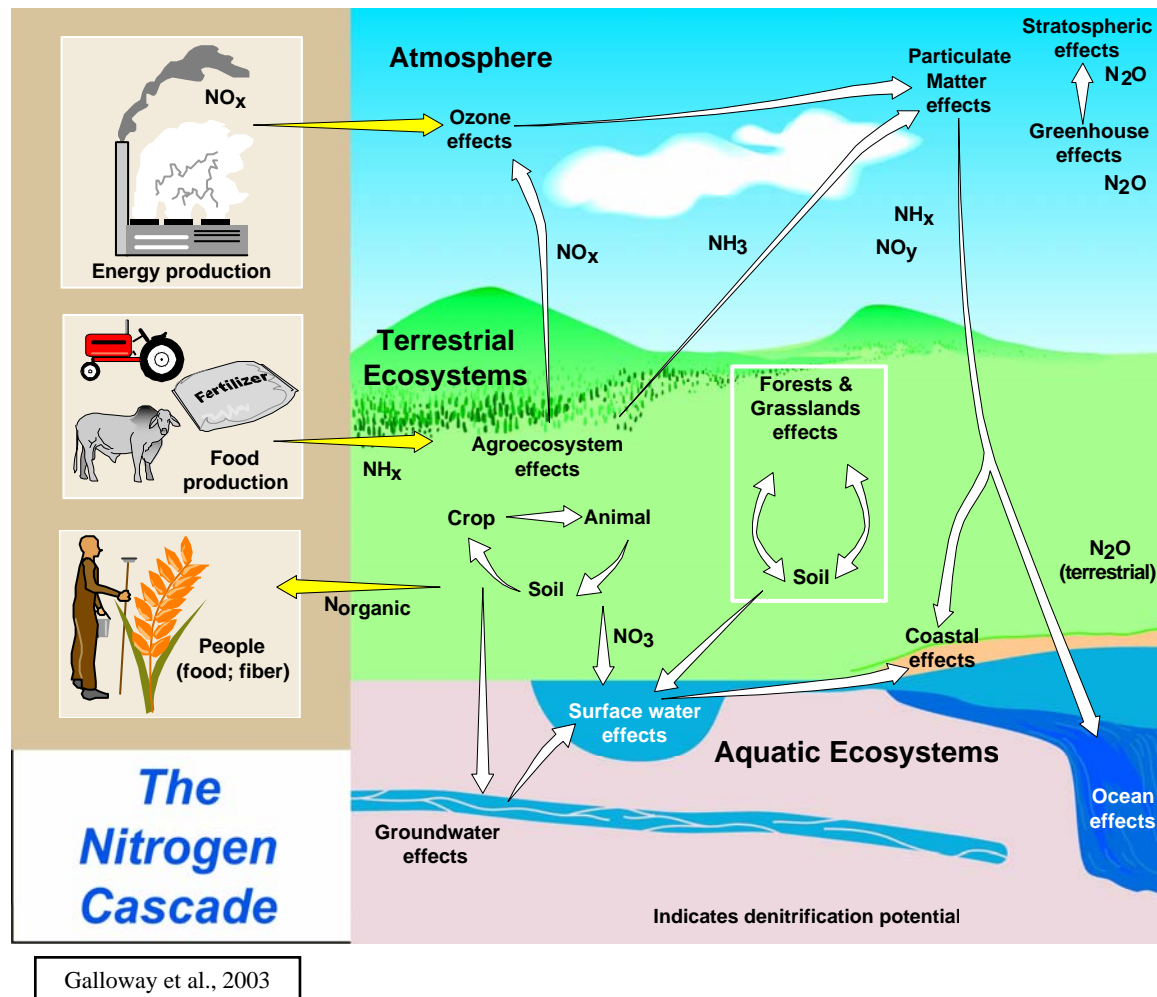
B.1.3.2 Ecosystem services relating to NO_x/SO_x secondary standard issues

Ecosystems provide services through a complex web of ecosystem processes that create and sustain the services. For example, if we are interested in how terrestrial and riparian ecosystems can help regulate nitrogen inputs to aquatic systems, we should consider all the nitrogen transformations and flow paths within these ecosystems (Figure B.3). Ecosystem processes can help mitigate adverse effects from environmental stressors, but processes can also be negatively affected by environmental stressors, decreasing the amount of ecosystem services they provide. For example, plant nutrient uptake and microbial denitrification and immobilization decrease the nutrient content in soil water, decreasing the potential for leaching into aquatic systems. Environmental stressors that decrease plant growth can increase nutrient flows to aquatic systems.

The web of ecosystem processes that influence the regulation of nitrogen transport to aquatic systems will also affect other ecosystem services such as carbon sequestration, so that

management activities that alter these processes will affect a bundle of services, not just one. The impact of a land-use change will not have the same impact on all of the ecosystem services; while one ecosystem service may be improved by a land-use change, it may come at a cost of decreasing another ecosystem service. Food production is a classic example of tradeoffs between ecosystem services. Adding fertilizer to crops will increase food production but may decrease water quality as some fertilizer ends up in the waterways. Thus, when ecosystems services are quantified, it is imperative that the entire bundle of services is evaluated, and that the tradeoffs between ecosystem services be included in the quantification.

Figure B.3. The Nitrogen Cascade illustrates the complexity of pools and processes at the watershed scale that are interconnected with the addition of nitrogen to the system, and with the output of the system in terms of ecological services to benefit man. Adapted from, Galloway et al. 2003 (Ecological Research Program Strategic Directions, ORD, April 2007, Preliminary Draft)



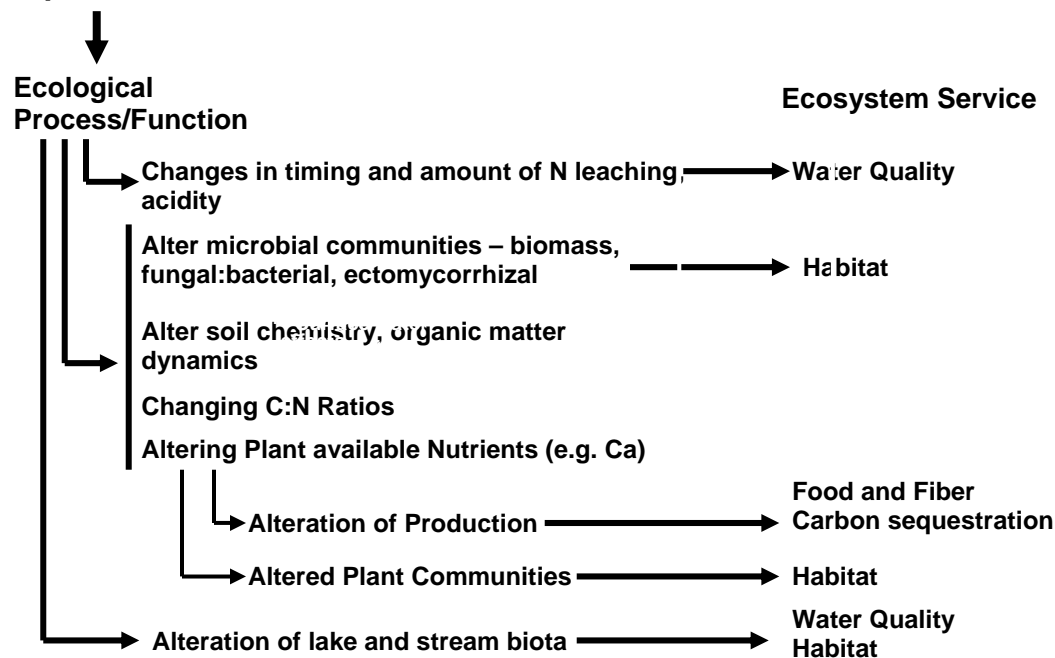
1 **B.1.3.3 Valuation**

2 While not the sole means of assessing adversity of effect, ecological improvements can
3 benefit people indirectly by increasing the delivery of “ecosystem services,” which are the end
4 products of ecological functions important to humans (Dailey 1997, Balmford et al. 2002, NRC
5 2004, Banzhaf and Boyd 2005, Boyd and Banzhaf 2006). Such valuable ecological functions
6 include the maintenance of stable climate conditions, the regulation of water availability and
7 quality, and nutrient retention (Dailey 1997). Through their effect on ecosystem services,
8 ecological improvements may lead to improved agricultural yields, recreational opportunities,
9 human health or other types of benefits. For example, protecting wetlands and the natural flow
10 regulation and water purification services they provide may lead to enhanced recreational fishing
11 or swimming opportunities in connected water bodies, reduced flooding in downstream
12 residential areas, or reduced incidences of illness from contaminated drinking water. Ecological
13 improvements may also benefit people directly through aesthetic improvements or through
14 increases in “nonuse” values (e.g., NRC 2004), which can arise from a variety of motivations
15 including an intrinsic concern for the existence of species populations or ecosystems in a
16 relatively undisturbed state or a desire to preserve healthy ecosystems for future generations.

17 Valuation is a tool for integrating information on the effects of NO_x and SO_x on the
18 multiple ecosystem processes that form the basis for the ecosystem service. It provides a
19 common currency to compare impacts. The National Research Council (2004) noted that the
20 translation from ecosystem structure and function to ecosystem goods and services is described
21 by an ecological production function (goods and services that can be produced from inputs of
22 natural and human capital, labor, and other resources). Many ecological relationships can be
23 quantified using existing data and models, at least for particular settings where data are
24 quantified and models have been calibrated. A very simple example of some ecological
25 processes or functions that would need to be quantified in order to determine the impact of
26 nitrogen deposition on an ecological service is shown in Figure B.4.

FigureB.4. Ecological Relationships Potentially disrupted by nitrogen deposition and the Ecosystem Service potentially affected

N Deposition to Watershed



A number of different approaches have been suggested for valuing ecosystem services:

1) Provisioning services commodity values. There are methods and data for obtaining monetary values for commodities under scenarios of change. These break down into ‘natural resource values’ for extracted goods (e.g. cut and sell the tree in the market) and ‘environmental quality values’ to leave the tree in the forest and obtain the benefits of it in its role in the ecosystem.

2) Other kinds of monetization. Economic methods that have been used to value ecological improvements include production or cost function approaches, travel cost models, hedonic property models, and stated preference surveys. Bioeconomic modeling, which involves combining models of species population or ecosystem dynamics with economic models of human behavior, is another approach that can potentially be used to value ecological improvements. Most bioeconomic models have been applied to fishery and forestry management problems, but in many cases these models could be adapted to estimate WTP for

1 environmental improvements that may affect the growth rates or carrying capacities of the focal
2 species.

3 3) When direct valuation is infeasible, benefits or value transfer methods may be
4 employed. This method takes direct values for ecosystem services from a primary study for a
5 particular location and transfers those values to a different location using either a point estimate
6 or value transfer function. Direct transfer of point value estimates has the potential to lead to
7 large uncertainties and in some cases biases in estimated values for ecosystem services, because
8 of the high dependency of most ecosystem values on the specific characteristics of the location in
9 which they were collected. Value transfer functions which allow the values to vary with
10 ecosystem characteristics can provide for more accurate estimation, but still have large
11 uncertainties.

12 4) Non-monetary values for individual and bundled services. These are developed from
13 experimental field studies, literature and modeling activities. There are two types of non-
14 monetary valuation methods, one based on direct physical or biological tradeoffs, and the other
15 based on tradeoffs defined by human preferences for the ecosystem functions defined by
16 different bundles of physical and biological attributes. Figure B.5 illustrates a conceptual
17 approach for developing non-monetary values using physical and/or biological tradeoff
18 functions. Ecological response functions (ERFs) and ecological tradeoff functions (ETFs) can be
19 developed to quantify the response of a service to changes in NO_x/SO_x concentrations and the
20 tradeoffs between different services given these ambient air quality drivers. ERFs and ETFs will
21 have to be, for the most part, defined by data mining from existing published literature. In the
22 future, however, these quantifying functions will be the focus of research to broaden the scope of
23 the science assessment and risk/exposure assessment. Process-based models will be used to (1)
24 synthesize/link the suite of ERFs and ETFs and (2) generate maps and summaries of ecosystem
25 services and tradeoffs in response to current and future ambient air indicators for NO_x and SO_x.
26 Ecological response functions are developed to show the response of a service to changes in an
27 ecosystem driver and ecological tradeoff functions combining response functions to show the
28 tradeoffs between different services given a driver. By describing the differing bundles of
29 ecosystem services under varying levels of NO_x and SO_x and offering choices between those
30 bundles, tradeoff or indifference curves can also be generated reflecting individual and
31 population level preferences for different ecosystem functions. These tradeoffs can be presented
32 to survey respondents using methods such as conjoint analysis, to provide measures of

September 2007 B-15 DRAFT-DO NOT QUOTE OR CITE

1 preferences for different levels of ecosystem functions as expressed through ecosystem service
 2 levels. Figure B.6 shows an example of how ecosystem services information might be presented
 3 to a survey respondent in a conjoint analysis framework.
 4 Figure B.5. Conceptual approach for quantifying Ecological Response Functions (ERF) and
 5 Ecological Tradeoff Functions (ETF) for ecosystem services in relation to different levels of
 6 NO_x or SO_x (forcing variables)

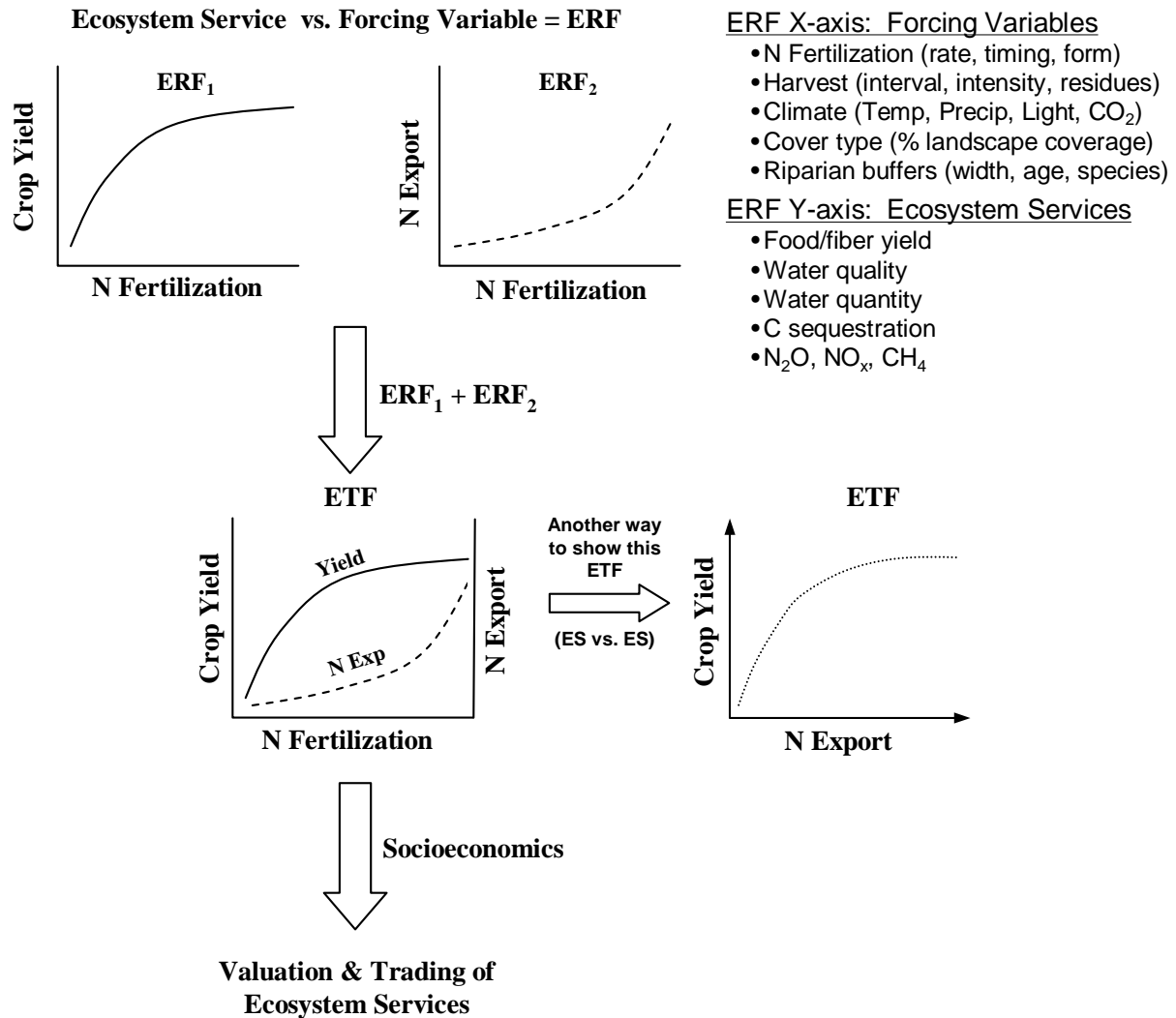


Figure B.6. Example Survey Design for Preference Based Ecosystem Tradeoffs

Example Conjoint Scenario Rating Form

Ecosystem Services Scenario Descriptions				
	Water Quality	Habitat	Food and Fiber Production	Carbon Sequestration
Scenario 1	Drinkable	Low diversity	High	High
Scenario 2	Drinkable	High diversity	Low	Low
Scenario 3	Swimmable	Medium diversity	Medium	Medium
Scenario 4	Boatable	Low diversity	High	Medium
Scenario 5	Swimmable	Low diversity	Low	Low

Scenario Rating									
Please rate how desirable each scenario is overall by circling one of the numbers in each row of the following table:									
	Highly Desirable	Quite Desirable	Desirable	Slightly Desirable	Neither Desirable nor Undesirable	Slightly Undesirable	Undesirable	Quite Undesirable	Highly Undesirable
Scenario 1	1	2	3	4	5	6	7	8	9
Scenario 2	1	2	3	4	5	6	7	8	9
Scenario 3	1	2	3	4	5	6	7	8	9
Scenario 4	1	2	3	4	5	6	7	8	9
Scenario 5	1	2	3	4	5	6	7	8	9